



THE UNIVERSITY *of* EDINBURGH

Edinburgh Research Explorer

Low coverage genomic data resolve the population divergence and gene flow history of an Australian rain forest fig wasp

Citation for published version:

Cooper, L, Bunnefeld, L, Hearn, J, Cook, JM, Lohse, K & Stone, G 2020, 'Low coverage genomic data resolve the population divergence and gene flow history of an Australian rain forest fig wasp', *Molecular Ecology*. <https://doi.org/10.1111/mec.15523>

Digital Object Identifier (DOI):

[10.1111/mec.15523](https://doi.org/10.1111/mec.15523)

Link:

[Link to publication record in Edinburgh Research Explorer](#)

Document Version:

Peer reviewed version

Published In:

Molecular Ecology

Publisher Rights Statement:

This is the peer reviewed version of the following article: Cooper, L., Bunnefeld, L., Hearn, J., Cook, J.M., Lohse, K. and Stone, G.N. (2020), Low coverage genomic data resolve the population divergence and gene flow history of an Australian rain forest fig wasp. *Mol Ecol*. Accepted Author Manuscript. doi:10.1111/mec.15523, which has been published in final form at <https://onlinelibrary.wiley.com/doi/10.1111/mec.15523>. This article may be used for non-commercial purposes in accordance with Wiley Terms and Conditions for Self-Archiving.

General rights

Copyright for the publications made accessible via the Edinburgh Research Explorer is retained by the author(s) and / or other copyright owners and it is a condition of accessing these publications that users recognise and abide by the legal requirements associated with these rights.

Take down policy

The University of Edinburgh has made every reasonable effort to ensure that Edinburgh Research Explorer content complies with UK legislation. If you believe that the public display of this file breaches copyright please contact openaccess@ed.ac.uk providing details, and we will remove access to the work immediately and investigate your claim.



Low coverage genomic data resolve the population divergence and gene flow history of an Australian rain forest fig wasp

Lisa Cooper^{1#}, Lynsey Bunnefeld^{2#}, Jack Hearn^{1,3}, James M Cook⁴, Konrad Lohse^{1*},
Graham N. Stone^{1*}.

1. Institute of Evolutionary Biology, University of Edinburgh, EH9 3FL Edinburgh, UK.

2. Biological & Environmental Sciences, University of Stirling FK9 4LA Stirling, UK.

3. Vector Biology Department, Liverpool School of Tropical Medicine, Liverpool, UK.

4. Hawkesbury Institute for the Environment, Hawkesbury Campus, Western Sydney

University, Richmond, New South Wales, Australia

Corresponding author: Prof. Graham N. Stone, Institute of Evolutionary Biology, Charlotte

Auerbach Road, University of Edinburgh, EH9 3FL Edinburgh, UK.

email: graham.stone@ed.ac.uk

these authors are joint first authors of the paper

* these authors contributed equally to this work.

Key words: Demography, phylogeography, Australia, *Ficus*, *Pleistodontes*, Agaonidae.

Running title: Population divergence and gene flow in a fig wasp

Orcid identifiers:

L. Bunnefeld: [0000-0002-9226-7153](https://orcid.org/0000-0002-9226-7153)

J.M. Cook: 0000-0001-8447-6126

J. Hearn: 0000-0003-3358-4949

K. Lohse: 0000-0001-9918-058X

G.N. Stone: 0000-0002-2737-696X

Abstract

Population divergence and gene flow are key processes in evolution and ecology. Model-based analysis of genome-wide datasets allows discrimination between alternative scenarios for these processes even in non-model taxa. We used two complementary approaches (one based on the blockwise site frequency spectrum (bSFS), the second on the Pairwise Sequentially Markovian Coalescent (PSMC)) to infer the divergence history of a fig wasp, *Pleistodontes nigriventris*. *Pleistodontes nigriventris* and its fig tree mutualist *Ficus watkinsiana* are restricted to rain forest patches along the eastern coast of Australia, and are separated into northern and southern populations by two dry forest corridors (the Burdekin and St. Lawrence Gaps). We generated whole genome sequence data for two haploid males per population and used the bSFS approach to infer the timing of divergence between northern and southern populations of *P. nigriventris*, and to discriminate between alternative isolation with migration (IM) and instantaneous admixture (ADM) models of post divergence gene flow. *Pleistodontes nigriventris* has low genetic diversity ($\pi = 0.0008$), to our knowledge one of the lowest estimates reported for a sexually reproducing arthropod. We find strongest support for an ADM model in which the two populations diverged *ca.* 196kya in the late Pleistocene, with almost 25% of northern lineages introduced from the south during an admixture event *ca.* 57kya. This divergence history is highly concordant with individual population demographics inferred from each pair of haploid males using PSMC. Our analysis illustrates the inferences possible with genome-level data for small population samples of tiny, non-model organisms and adds to a growing body of knowledge on the population structure of Australian rain forest taxa.

50 Introduction

Division of an ancestral population into daughter populations is a universal and
52 repeating process in biology. The tempo and mode of population divergence are central to
evolutionary processes ranging from local adaptation and range expansion to the origin of
54 species (Hey & Nielsen 2004; Martin et al., 2013; Sousa & Hey, 2013). From a
demographic perspective, population divergence can be described in terms of the sizes of
56 ancestral and descendant populations, the time at which the ancestral population split, and
parameters capturing the timing, extent and direction of gene flow between descendant
58 populations. Gene flow can be modelled in at least two general ways (Fig. 1): an isolation
with migration (IM) model identifies gene flow as the result of ongoing dispersal (Nielsen &
60 Wakeley 2001, Hey, & Nielsen 2004, Lohse et al 2011), while an instantaneous admixture
(ADM) model associates gene flow with one or more discrete dispersal events in the past
62 (Durand et al., 2011; Sousa & Hey, 2013; Lohse & Frantz, 2014). Thus, which of these
models applies may tell us whether putative dispersal barriers are past or ongoing, and
64 help to identify evolutionarily independent conservation units within species. The models
also have very different implications for local adaptation: while local adaptation may
66 proceed unimpeded during periods of complete isolation in the ADM model, continuous
gene flow under the IM model imposes a constant genetic load of locally deleterious
68 variants (Bisschop et al 2020).

Discriminating among alternative models of gene flow and estimating the relevant
70 demographic parameters are data-hungry problems and a growing number of approaches
exploit the signal contained in whole genome data (WGD) for a small sample of individuals
72 (Li & Durbin, 2011; Lohse & Frantz 2014; Bunnefeld et al., 2018). Whilst these approaches
are currently applicable to a limited diversity and complexity of demographic models
74 (discussed further below), their requirement for only a small number of relatively low
quality genomes makes them accessible and affordable for non-model taxa, including rare

76 taxa for which larger samples of individuals and reference genomes often do not exist
(Allendorf et al., 2010; Fuentes-Pardo & Ruzzante 2017).

78 Here we use two such approaches to infer the population divergence history of
Pleistodontes nigriventris, a wasp that is the only pollinator of an endemic fig species
80 (*Ficus watkinsiana*) found in two widely separated blocks of rainforest along the east coast
of Australia (see below) (Dixon 2003; Lopez-Vaamonde et al., 2002). Our overall objective
82 is to infer the extent and direction of fig wasp gene flow between these two populations,
using WGD for just two individuals per population. Our approaches take advantage of the
84 haplodiploidy of the Hymenoptera, by sampling males whose haploid genomes facilitate
data analysis and interpretation.

86 We first compare support for alternative IM and ADM models using a parametric
maximum-composite likelihood method (Lohse et al., 2011; 2016), and then compare
88 these results with those obtained using the Pairwise Sequentially Markovian Coalescent
(PSMC) (Li & Durbin, 2011), a non-parametric method. Both are well-suited to the pairwise
90 population divergence hypotheses we explore in *Pleistodontes nigriventris*. We chose
these methods because they infer population history based on different aspects of
92 genome-wide sequence variation, and have contrasting limitations. The composite
likelihood framework developed by Lohse et al. (2016) is based on a blockwise summary
94 of sequence variation, while PSMC exploits the information contained in the density of
pairwise differences along a minimal sample of two haploid genomes. The blockwise
96 method – by design – lacks power to detect very gradual demographic changes (for
example, in population size), while PSMC is known to smooth out very sudden changes (Li
98 & Durbin, 2011). The two methods therefore complement each other and together provide
a comprehensive picture of population history.

100 Fig trees are keystone biological resources in tropical and subtropical habitats
worldwide (Cook & Rasplus 2003; Harrison 2005), and fig fruits and their insect inhabitants

are important model systems in the study of community assembly and coevolution (Cook & Rasplus 2003; Segar et al., 2014). Our target species *Pleistodontes nigriventris* is the specialist pollinating wasp of *Ficus watkinsiana* (Lopez-Vaamonde et al., 2002; Male & Roberts 2005, Rønsted et al., 2008), a monoecious rainforest fig restricted to northern and southern populations over 1000km apart along the eastern coast of Australia (Fig. 2) (Dixon 2003). While the existence of intervening *F. watkinsiana* trees and associated *P. nigriventris* cannot be categorically excluded, no populations are known; the demographic history of *P. nigriventris* can thus be modelled in terms of pairwise population divergence.

The distribution of Australian rainforests has been dictated by two major processes. First, falling temperatures during the Miocene restricted them to areas of higher rainfall along the eastern and southern coasts (Markgraf et al., 1995), separated from drier inland habitats by the Great Dividing Range (Chapple et al., 2011) (Fig. 2). Second, during the Pleistocene climate oscillations, east coast Australian rain forests repeatedly expanded from, and contracted into, a latitudinal series of refugia separated by intervening areas of dry forest and shrublands (Bryant & Krosch, 2016; Chapple et al., 2011). The rainforest areas occupied by *F. watkinsiana* are currently separated by two major dryland corridors (Fig. 2): the Burdekin Gap, located between Mackay and Townsville, is the largest dry land corridor on the east coast, and the St. Lawrence Gap is a smaller lowland dry corridor located 350km further south (Weber et al., 2014; Bryant & Krosch, 2016). While the formation and stability of these dryland corridors through time is incompletely characterised (Bryant & Krosch, 2016), both have been implicated in restricting dispersal and driving population divergence in rainforest plants (Burke et al., 2013), including *F. watkinsiana* (Dixon, 2003H; Haine & Cook, 2005), and animals (e.g. Schauble & Moritz, 2001; Pope et al., 2001; Nicholls & Austin, 2005; Brown et al., 2006; Dolman & Moritz, 2006; Baker et al., 2008, MacQueen et al., 2012; Rix & Harvey, 2012; Bryant & Fuller, 2014; Bryant & Krosch, 2016).

128 The impact of biogeographic barriers on the genetic makeup of any species
depends on how long ago population divergence occurred, the sizes of the populations,
130 and the direction, mode and frequency of gene flow between them (Aeschbacher et al.,
2017; Ringbauer et al., 2018). Previous studies of fig/fig wasp systems provide evidence
132 for two contrasting paradigms with which patterns in *Pleistodontes nigriventris* can be
compared. Though female fig wasps are poor active flyers (and the males are wingless
134 and do not leave their natal fig) (Ware & Compton 1994a,b), they can be dispersed over
large distances by wind currents (Ahmed et al., 2009; Liu et al., 2015), particularly when
136 emerging from fig fruits high in the forest canopy (Harrison & Rasplus, 2006; Kobmoo et
al., 2010; Yang et al., 2015; Liu et al., 2015; Sutton et al., 2016). Long range dispersal by
138 fig wasps is supported by lack of genetic structure over distances ranging from several
hundred to >1500 km in pollinating fig wasps (Kobmoo et al., 2010; Liu et al., 2015; Tian et
140 al. 2015; Bain et al., 2016) and in host figs - particularly monoecious species that are often
large trees (Nazareno et al., 2013; Bain et al., 2016). *Ficus watkinsiana* can grow to 50m,
142 and if *Pleistodontes nigriventris* benefits from wind-assisted dispersal, we might expect to
find significant gene flow between *F. watkinsiana* populations. In contrast to this pattern,
144 local adaptation of diverging populations to local host figs and/or abiotic conditions could
drive genetic divergence between fig wasp populations over similar or even relatively small
146 geographic scales. Genetic divergence has been documented over tens of km between
mainland and island populations of a Chinese fig wasp (Tian et al., 2015), and between
148 the same northern and southern rainforest habitats studied here for two *Pleistodontes*
pollinators of another monoecious fig, *Ficus rubiginosa* (Darwell et al., 2014).

150 Here we answer the following questions for *Pleistodontes nigriventris*:

- 1) Is there evidence of genetic divergence between populations either side of the Burdekin
152 and St. Lawrence Gaps?
- 2) Is there a signal of post-divergence gene flow, and if so, in which direction?

- 154 3) Can we discriminate between the IM and ADM models of gene flow?
- 4) Do the blockwise and PSMC methods infer concordant population histories?
- 156 5) Is the inferred divergence time for *P. nigriventris* concordant with estimates for other co-distributed taxa?

158

Materials and Methods

160 Sample Collection

Samples were collected between January 2001 and August 2009 from four sites in Queensland, two in each of the northern and southern ranges of *F. watkinsiana*. The two Northern individuals were sampled from Kairi (N1: 17.21° S, 145.55° E) and Kamerunga (N2: 16.87° S, 145.68° E) and the two South individuals were sampled from Settlers Rise (S1: 27.68° S, 153.26° E) and Main Range (S2: 28.07° S, 152.41° E) (Fig. 2). Near-to-ripe *F. watkinsiana* fruits were collected and placed individually into specimen pots. Our sampling targeted male fig wasps, whose haploid genome facilitates analysis (Bunnefeld et al., 2018; Hearn et al., 2014). Once the wasps started to emerge (12-24 hours after collection depending on fig ripeness) figs were dissected and live males were placed directly into 70% ethanol to preserve for DNA extraction. Due to potentially high levels of sib-mating in fig wasps (Greeff et al., 2009; Sutton et al., 2016) all individuals were sampled from different figs.

174 DNA extraction and sequenced-based confirmation of identity.

DNA was extracted from whole male wasps 1.5-1.6mm long (Lopez-Vaamonde et al., 2002) using the Qiagen DNeasy Blood and Tissue Extraction kit. The Purification of Total DNA from Animal Tissues (Spin-Column) Protocol was followed with the following modifications to maximise DNA yield from these extremely small wasps. Step 1: Individual wasps were placed in 180 µl of buffer ATL and crushed using a mini-pestle. Step 2:

180 Riboshredder RNase (Epicentre) was used in place of RNase A. Steps 7-8: Buffer EB was
used in place of Buffer AE for the elutions as Buffer AE contains EDTA, which will interfere
182 with the downstream library preparation. Extractions were eluted in smaller volumes (25 µl)
than recommended and samples were incubated for longer (5 minutes). This protocol
184 yielded 35.7-58.3 ng of DNA per wasp, despite their very small size (total body length *ca.*
1mm).

186 Identification of male fig wasps was confirmed by comparison of sample sequences
to voucher sequences for a 433 base pair (bp) fragment of the mitochondrial cytochrome b
188 (cytb) gene (Lopez-Vaamonde et al., 2001). Sequences were amplified using the primers
CB1/CB2 (Jermiin & Crozier, 1994). 0.3 µl of DNA extraction was used per PCR reaction.
190 The remainder of the PCR mix consisted of 2 µl BSA (10 mg/ml), 2 µl 10X PCR buffer, 0.8
µl MgCl₂ (50mM), 0.3 µl of each primer (20 µM), 0.16 µl dNTPs (each 25 mM) and 0.1 µl
192 Taq (Bioline 5U/µl), made up to 20 µl with autoclaved MilliQ water. Amplification was
carried out using a Bio-Rad S1000 thermal cycler for 2 minutes at 94°C, 35 cycles of 30
194 seconds at 94°C, 30 seconds at 48°C, 40 seconds at 72°C, and a final elongation step of 5
minutes at 72°C. PCR products were visualised on a 2% agarose gel and cleaned using a
196 shrimp alkaline phosphatase and exonuclease 1 protocol. 2.5 µl of SAPExo1 mix (1.425 µl
SAP dilution buffer, 1 µl SAP (1U), 0.075 µl Exo1 (1.5U)) was added to each sample
198 before being incubated for 40 minutes at 37°C followed by 15 minutes at 94°C. Only the
forward strand was sequenced for each individual, using BigDye chemistry on an ABI 3730
200 machine at the Edinburgh Genomics facility. Sequences for the four individuals have been
deposited on GenBank (individual accession numbers: N1 MF597824; N2 MF597800; S1
202 MF597825; S2 MF597826).

204 **High-throughput library preparation and sequencing**

We generated an Illumina Nextera genomic library for each individual male fig wasp following the manufacturers' instructions. To make best use of paired end sequencing, the library fragment size distribution should be unimodal with a majority of fragments longer than the 150bp read length. We checked the fragment size distribution by running 1µl of each library on a high sensitivity DNA Bioanalyzer chip (Agilent 2100). Libraries for each individual were end-labelled using a unique pair of indices, pooled and sequenced in one lane of 150bp paired-end reads on the Illumina HiSeq platform at the Edinburgh Genomics facility. Pooling volumes were calculated to achieve ~6-fold coverage per individual assuming 50 gigabases of data per lane and a genome size of 300 Mb based on an estimate for another Agaonid fig wasp, *Ceratosolen solmsi* (278 Mb; Xiao et al., 2013). Our strategy in balancing sequencing depth across individuals was to maximise data availability for our analyses while minimising cost, and to avoid biases in variant calling that could result from sequencing one or more individuals to substantially greater depth. Final coverage for each individual matched expectations and is shown in Table S2. The short read data have been deposited at the ENA short read archive (Cooper et al., 2020a).

220

Bioinformatic pipeline

Reads for all four individuals were screened to remove low quality and contaminant reads and combined. After exclusion of contaminants, reads for all four individuals were combined to generate a *de novo* meta-assembly for *P. nigriventris* using *SPAdes* (version 3.6.2) (Bankevich et al., 2012). Reads from each individual were mapped back to this reference using BWA; variants were called using *GATK* (Van der Auwera et al., 2013; version 3.5.0) haplotype caller. After masking repeat sequences, sites with a minimum coverage of two and a mapping quality (>20) and base quality (>10) were identified using the *GATK* tool *CallableLoci*. The bioinformatic pipeline is summarised diagrammatically in Supplementary information, Figure S1.

(i) *Quality control and processing of sequencing reads*. To exclude low quality sequence,
232 reads were checked using *Fast-QC* (www.bioinformatics.babraham.ac.uk/projects/fastqc)
and trimmed at a base quality score of 20 (sliding window 15:20 and min length 50).
234 Adapters were trimmed using *Trimmomatic* version 0.32 (Bolgar et al., 2014). Any
adaptors found to be present after a second pass through *Fast-QC* were removed using
236 *Cutadapt* (Martin, 2011). Paired-end reads were merged (assembled into pairs) using
PEAR version 0.9.0 (Zhang et al., 2014) for *de novo* assembly only.

(ii) *Filtering of contaminant sequences*. Genomic data obtained from whole organism
libraries commonly contain reads from associated non-target organisms, including
240 symbionts, parasites and commensals. *Wolbachia* bacteria are common endosymbionts in
fig wasps (Haine and Cook, 2005) and diverse fungal taxa have been identified from a
242 genome assembly of the fig wasp *Ceratosolen solmsi* (Niu et al., 2015). To remove
contaminant sequences from our data, we first aligned reads across all 4 individuals to
244 create a single *Velvet* reference assembly (version 1.2.10) with a k-mer length of 31
(Zerbino and Birney, 2008). The filtered reads for each individual were then mapped to this
246 *Velvet* reference assembly using *Bowtie2* version 2.2.3 (Langmead and Salzberg, 2012).
Contaminant sequences were identified using blobtools (Kumar et al., 2013), which uses
248 BLAST searches to create Taxon-Annotated-GC-Coverage plots (Blobplots) that allocate
aligned reads to taxa at a user-specified level. We used four BLAST approaches to assess
250 taxonomic matches: (i) The fast protein aligner *Diamond* (version 0.7.9) (Buchfink et al.,
2015) was used alongside three BLAST (version 2.2.29) searches: (ii) against the NCBI
252 nucleotide database (<https://www.ncbi.nlm.nih.gov/nucleotide/>), (iii) against the genome of
the Agaonid pollinating fig wasp, *Ceratosolen solmsi* (Xiao et al., 2013), and (iv) against
254 the genome of the pteromalid parasitoid wasp *Nasonia vitripennis* (Werren et al., 2010).
Non-Arthropod reads (primarily allocated to Proteobacteria and Ascomycota) were
256 excluded from further analysis.

(iii) *Generation of a P. nigriventris reference assembly.* After exclusion of contaminants, reads for all four individuals were combined and re-assembled using *SPAdes* (version 3.6.2) (Bankevich et al., 2012). The quality of this reference assembly was assessed using *BUSCO* (*Benchmarking Universal Single-Copy Orthologs*) version 1.1b1 (Simão et al., 2015) using the Arthropoda *BUSCO* set (<http://busco.ezlab.org>).

(iv) *Variant calling.* Filtered, merged reads from each individual were mapped back to the reference meta-assembly using the *Burrows-Wheeler Aligner (BWA)* version 0.7.10 (Li & Durbin, 2009). FASTA file indexes and sequence dictionaries were created using *samtools* (Li et al., 2009) *faidx* (version 1.2) and the *picard* (<https://broadinstitute.github.io/picard/index.html>) tool *CreateSequenceDictionary* (version 1.141) respectively. The BAM files were sorted and merged to create a single species BAM file using the *picard* tool *MergeSamFiles*. Duplicate reads were removed from these merged files using the *picard* tool *MarkDuplicates*. Variants were called using *HaplotypeCaller* in *GATK*. As male wasps are haploid, ploidy was set to 1 and the ‘-emit variants only’ option was used. Only SNPs were considered.

(v) *Masking repetitive regions.* Regions of repetitive DNA can cause assembly and mapping errors in short read data (Treangen & Salzberg, 2011). We created a library of repetitive regions in the reference meta-assembly using *RepeatScout* (version 1.0.5) (Price et al., 2005), and masked these using *RepeatMasker* (version open-4.0.6) (Smit, AFA, Hubley, R & Green, P. *RepeatMasker Open-4.0*. 2013-2015 <http://www.repeatmasker.org>). *RepeatMasker* outputs an annotation file which was used to create a BED file of repeat positions for downstream processing.

(vi) *VCF filtering.* High quality sites were identified based on coverage and mapping quality using the *GATK* tool *CallableLoci* and default parameter values except for the following: minimum base quality of 10, minimum mapping quality of 20, minimum read depth of two reads per site. The base quality score recalibrated (BQSR) BAM file was subsampled to

extract BAM files for each individual. We used *CallableLoci* to generate BED files of
284 callable regions in each individual. Repeat regions were excluded and positions meeting
filters across all four individuals extracted using *bedtools multiIntersectBed*. This BED file
286 was used to filter the VCF files (generated from the *GATK* pipeline) for callable variable
sites using *bcftools* (Li, 2011) version 1.2.

Fitting of IM and ADM models

290 (i) *Specification of alternative IM and ADM divergence models*: Both the IM and the ADM
model assume an ancestral population with a constant effective population size (N_a) that
292 splits into two populations (North and South). One of the descendant populations
maintains the same ancestral population size N_a , whilst the other is free to change to $(1/B)$
294 $\times N_a$ (i.e. B scales the rate of coalescence in the other population). In both models (Fig. 1),
population divergence occurs at time T in the past. In the IM model divergence is followed
296 by continuous unidirectional migration at rate $M = 4Nm$ migrants per generation (m is the
per lineage probability of migrating). In the ADM model (Fig. 1), gene flow occurs through
298 an instantaneous and unidirectional admixture event at time T_{adm} which transfers a fraction
 f of lineages from the donor population into the recipient (Fig. 1). All time parameters are
300 scaled in $2N_a$ generations. We assessed support for all four possible combinations of
population size and gene flow under both the IM and ADM models (Fig. S2), as well as for
302 simpler nested models which either involving gene flow but a single population size
parameter ($B=1$) or two population sizes but no post-divergence gene flow.

304 (i) *Generation of blockwise site frequency spectra (bSFS)*: We used the composite
likelihood calculation described in Lohse et al. (2016) to fit the IM and ADM models (see
306 also Jordan et al., 2017 and Nürnberg et al., 2017). The method uses information in
patterns of linked sequence variation contained in short blocks and has previously been
308 used for demographic inference under a variety of demographic scenarios including the IM

model (Lohse et al., 2012) and models of discrete admixture (Bunnefeld et al., 2018). For
310 a sample of four haploid individuals (two from each of the Northern and Southern
populations), we can distinguish four mutation types: variants in the two Northern samples
312 (Nvar), variants in the two Southern samples (Svar), fixed differences between North and
South (Fixed differences), or variants shared between North and South (Sharedvar)
314 (Lohse et al., 2016). The site frequency spectrum of a block (bSFS) consists of counts of
these four site types, i.e. vector {Nvar, Svar, Sharedvar, Fixed}. We used the automated
316 recursion implemented by Lohse et al. (2016) to obtain the generating function of
genealogies under the ADM model and computed the composite likelihood of both the
318 ADM and IM models in *Mathematica* as described in Lohse et al. (2016) and Nürnberger et
al. (2017).

320 The likelihood calculation assumes an infinite sites mutation model and no within-
block recombination. Given these assumptions, blocks that contain both fixed difference
322 and shared variants (violating the 4-gamete test) are not possible and were excluded from
the likelihood calculation. We chose a block length that strikes a balance between potential
324 bias (arising from recombination within blocks) and power (which suffers when too few
blocks contain multiple variant sites). We used a custom python script to extract aligned
326 sequence blocks of a fixed length of 387 base pairs, which corresponds to an average of
1.5 variant sites per block (see Table S1). The blockwise data were analysed in
328 *Mathematica* (Wolfram Research, Inc., Mathematica, Version 10.4, Champaign, IL (2016))
(Cooper et al., 2020b). The computational cost of calculating composite likelihoods
330 increases with the number of unique bSFS configurations considered in the data. We
limited the number of bSFS configurations by lumping mutation counts above a threshold
332 $k_{\max} = 2$ for the Nvar, Svar and SharedVar and $k_{\max} = 3$ for fixed differences. Since 87% of
the blockwise data are within these k_{\max} bounds and so included in the composite
334 likelihood calculation exactly, the expected loss of power is minimal.

(iii) *Estimation of demographic parameters*: We estimated all model parameters for both
336 the IM and ADM model. Estimation of N_a (as $N_a = \theta / (4 \mu)$) and scaling of T parameters
into years requires an estimate of the mutation rate per generation (μ) and the number of
338 fig wasp generations (g) per year. In the absence of any fig wasp estimate of μ , we used
the per generation mutation rate of 2.8×10^{-9} for *Drosophila melanogaster* (Keightley et al.,
340 2014). Note that the relative timing of events in different models is not affected by this
calibration. We assumed four fig wasp generations per year ($g=4$) and converted time
342 estimates into years by multiplying by $2N_a$ and dividing by g . The generation time for *P.*
nigriventris is not known with certainty, and we assume 4 (range of 2-6) generations per
344 year based on the following rationale. Duration of any single fig wasp generation can be
estimated by measuring the time from pollination to ripening of individual figs. This is
346 because wasp eggs are laid at the time of pollination, the emerging daughters disperse at
the same time as the fig ripens, and these new females lay eggs rapidly as they live only a
348 day or two as adults (Sutton et al. 2018). Estimates of the length of fig tree reproductive
events are available for several species (e.g. Bronstein 1990), but not available for *P.*
350 *nigriventris*. Jia et al. (2008) estimated a generation time of about 40 days for another
Australian *Pleistodontes* fig wasp (*P. imperialis*) pollinating *Ficus rubiginosa*. This might
352 suggest 8-9 generations per year for *P. imperialis*, but while these monoecious figs
(including *F. watkinsiana*) fruit asynchronously year-round (Harrison, 2005), they are in
354 highly seasonal environments and generation time can be much longer in cooler or drier
periods. Fruit development time is also longer in species producing larger figs. Fruit
356 development takes 3-8 months in winter figs of *F. macrophylla* (JMC, unpublished data),
which has smaller figs than *F. watkinsiana* (Al-beidh, 2010). Hence, a range of 2-6
358 generations per year seems likely for *P. nigriventris*. We discuss the consequences of
uncertainty in our generation time estimates.

360 (iv) *Correcting for the effects of linkage between blocks.* The composite likelihood
calculations do not account for linkage between blocks. Given the difficulty of assessing
362 the level of linkage between blocks in highly fragmented assemblies, we incorporated
linkage effects using a parametric bootstrap approach (Lohse et al., 2016). We used
364 *msprime* (version 0.3.1 (IM) and 0.4.0 (ADM)) (Kelleher et al., 2016) to simulate 100
datasets assuming a recombination rate of 2.719×10^{-10} per base (estimated from the
366 data, see below). Each simulated dataset had the same total length as the real data. To
keep simulations computationally tractable, we partitioned each simulated genome into 20
368 pseudo-chromosomes of equal length. Because our simulations allow for recombination
anywhere along a linear genome, this parametric bootstrap scheme allows us to quantify
370 both the uncertainty in estimates due to linkage between blocks and the potential bias due
to recombination within them. Fitting the inferred model to these simulated datasets, we
372 obtained 95% confidence intervals for parameter estimates as ± 1.96 standard deviations
(of the analogous estimates on the simulated data).

374 (v) *Model selection.* We used an analogous parametric bootstrap scheme to determine
whether the IM or ADM models provided a significantly better fit to our data. We fitted both
376 models to each of 100 bootstrap datasets simulated under the best-fit IM model
parameters. The distribution of differences in log composite likelihood (calculated as $\Delta \ln \text{CL}$
378 $= 2 * (\text{ADM} \ln \text{CL} - \text{IM} \ln \text{CL})$) between the IM model and the ADM model represents an
expectation of selecting the ADM when the IM is true (Type 1 error). We determined
380 critical values from this distribution at a significance level of 0.05.

(vi) *Recombination rate estimation.* There are no estimates of either the per generation
382 recombination rate (r) or the population recombination rate ($\rho = 4N_e r$) for any fig wasp
species. We therefore estimated a recombination rate for *P. nigriventris* using a two-locus
384 generating function that co-estimates ρ (scaled by $2N_e$) and θ , for a sample of two haploid

genomes drawn from a single panmictic population (here, the two Northern individuals)
(Lohse et al., 2011).

PSMC analysis

We compared the results of our blockwise analysis with inference using the Pairwise Sequentially Markovian Coalescent (PSMC) (Li & Durbin, 2011). PSMC infers the effective population size history from a pair of genomes, which may either be sampled from a single diploid individual, or a pair of haploid individuals. PSMC provides a continuous estimate of demographic parameters through time. While PSMC does not allow direct inference of population divergence, comparison of the PSMC trajectories of two diverged populations (here, North and South) reveals when these shared the same population size and so were likely part of the same ancestral population.

PSMC uses the density of pairwise differences along the genome to infer a trajectory of population size through time. Input files were generated from the per-individual BAM files using *samtools mpileup* (version 0.1.19). The pileup file was converted to a VCF file in *bcftools* (version 0.1.19) and a consensus sequence (fastq file) per individual was generated using the *bcftools* utility *vcf2fq*. Consensus files contained 31,371 and 33,523 contigs for the Northern and Southern pairs respectively. The two consensus files per population were merged using the *seqtk* function *mergefa* (version 1.0) and converted to the PSMC input format using the PSMC function *fq2psmcfa* (version 0.6.5). Only contigs that contain >10kb of unfiltered bases, $\geq 80\%$ of which pass a quality threshold score, here set to ≥ 20 were included (6,900 and 7,305 contigs for the North and South pairs respectively). We analysed each pair of individuals with PSMC (version 0.6.5) using the following parameter values: $N = 30$, $t = 20$, $r = 10.3$ and $p = "4+60*1+4"$. Each dataset was sub-sampled 100 times to generate bootstrap replicates using the PSMC

410 utility *splitfa* (version 0.6.5). Results were calibrated using the same mutation rate and generation times given above.

412

Results

414 **De novo genome assembly for *Pleistodontes nigriventris***

Nextera libraries were generated successfully for all four male *Pleistodontes*
416 *nigriventris* (Figure S3). After initial filtering and trimming the joint assembly of reads for all four individuals in *Velvet* contained 173,180 contigs ≥ 200 bp (Table S1). We identified
418 contaminant reads from a range of non-Arthropod taxa (Figure S4). Individual assemblies contained between 5k and 149k reads identified as bacterial. For three individuals over
420 97% of these were attributed to *Wolbachia* (and 24% in the final individual) (Table S2). The relative fraction of *Wolbachia* reads was significantly higher in Northern compared to
422 Southern individuals (Chi-squared = 95159, df = 1, $p < 2.2e-16$), consistent with previous findings of between-population differences in *Wolbachia* prevalence in *P. nigriventris*
424 (Haine & Cook, 2005). Excluding contaminant reads resulted in between 11 and 22.5 million reads per individual (Table S1). Re-assembly using *SPAdes* (Cooper et al., 2020a)
426 improved the contiguity (139,731 contigs ≥ 200 bp, N50 of 9,643 bp) and the completeness of the final assembly: *CEGMA* scores: 93.15% complete and 99.19% partially complete,
428 *BUSCO* scores for the reference Arthropod gene set: 74% complete, 4.5% duplicated, 16% fragmented and 8.4% missing. We note that the *P. nigriventris* assembly, although
430 fragmented, has higher completeness (in terms of complete Core Eukaryote Genes) than the genome of *Ceratosolen solmsi* (*CEGMA*: 88% complete), the only published fig wasp
432 genome (Xiao et al., 2013). Repetitive elements made up 21.4% of the reference assembly and were excluded from subsequent analyses.

434 Processing of the filtered and aligned reads (mean coverage per individual of 3.7x) identified 1,837,396 SNPs. The blockwise site frequency spectrum contained the four site

436 types in the following proportions: Nvar 0.164, Svar 0.153, Fixed 0.673, and Sharedvar
0.0102. The average pairwise diversity per site (after filtering) for the North and South
438 populations were very similar ($\pi = 0.000884$ and 0.000822 respectively). This is one of the
lowest estimates of genetic diversity reported for any sexually reproducing arthropod
440 (Leffler et al., 2012; Romiguier et al., 2014). Pairwise F_{st} was 0.74, indicating high
differentiation between the Northern and Southern populations.

442

Inferring divergence with continuous migration and admixture

444 Our block extraction protocol (Cooper et al., 2020b) resulted in 775,977 blocks of
387 base pairs. Of these, 0.8% (5,872) contained both shared variants and fixed
446 difference, violating the 4-gamete test, and were excluded from blockwise analyses.

Isolation with continuous migration (IM). Models incorporating post-divergence migration
448 and different effective population sizes (N_e) received substantially greater support than
otherwise equivalent models with no migration and/or a single N_e parameter (Table 1a).
450 The best-supported direction of migration is from South to North, irrespective of whether
we assumed a single N_e (compare model IM3 to IM2) or two N_e parameters (compare
452 IM7+IM9 to models IM6+IM8). The scenario in which the South population retained the
ancestral N_e (IM9; Figure 3a) had highest support. Under this model, the split between the
454 two populations occurred 177 (95% C.I. 172–182) thousand years ago (kya) and the N_e for
the ancestral/South population was estimated to be slightly higher (69k, 95% C.I. 65.4k-
456 72.6k) than for the North population (58k 95% C.I. 54.3k- 61.5k). We inferred a migration
rate M from south to north of 0.071. Although low (1 migrant every 28 generations), this
458 estimate was significantly greater than zero (95% C.I. 0.045 – 0.097).

Isolation with instantaneous admixture (ADM). Inferences for models assuming
460 instantaneous admixture mirrored those for IM models both in terms of model comparisons
and parameter estimates: scenarios with gene flow and two N_e parameters were

462 significantly better supported than otherwise equivalent models without admixture and/or a
single N_e (Table 1b). The best-supported admixture direction was again from South to
464 North, and the model assuming that the Southern population retained the ancestral N_e
(ADM10; Fig. 3a) had greatest support. N_e estimates under the best fitting ADM history
466 were similar to those under the analogous IM model (IM9): we inferred an
ancestral/Southern N_e of 69k (95% C.I. 66.1k-72.6k) and a lower Northern N_e of 59k
468 (55.2k-62.2k). In contrast, under the best fitting ADM9 the population split was estimated
196kya (95% C.I. 193-198kya), slightly older than under the analogous IM model. The
470 admixture event was inferred to have occurred 57kya (95% C.I. 53-62kya) (Table 1b) and
around a quarter (admixture fraction $f = 0.239$, 95% C.I. 0.206-0.272) of Northern lineages
472 are inferred to have originated from the Southern population.

Greater support of instantaneous admixture. The ADM model had greater support than the
474 IM model ($\Delta\ln\text{CL} = 4175$). Since the blockwise composite likelihood calculation ignores
linkage between blocks and given that IM and ADM models are not nested, we cannot use
476 a likelihood ratio test to compare the support for these models. To confirm whether the
simpler IM model fits significantly worse than the ADM model we obtained a critical value
478 for $\Delta\ln\text{CL}$ (155.4 at $p=0.05$) using a fully parametric bootstrap. The difference in model
support in favour of the ADM model $\Delta\ln\text{CL} = 4175$ in the real data far exceeds this and
480 confirms that admixture provides a better fit to our data than continuous gene flow. To
investigate which aspect of blockwise variation in the data allows discrimination between
482 instantaneous admixture (ADM) and continuous migration (IM), we compared the
frequencies of the most common bSFS configurations in the data with those expected
484 under the best fitting IM and ADM models (Figure S6). Inspection of the residual reveals
that the ADM model predicts both the frequency of monomorphic blocks ($\{0,0,0,0\}$) and
486 blocks with more than three fixed differences ($\{0,0,>3,0\}$) better than the IM model.

PSMC supports divergence and instantaneous admixture from South to North

We used PSMC (Li & Durbin 2011) to infer trajectories of N_e change for Northern and Southern populations. Comparing the trajectories both with each other and with parameters inferred under the best fitting models of divergence and gene flow (ADM10) reveals a close correspondence between blockwise analyses and PSMC in several respects (Fig. 4). First, PSMC trajectories are consistent with the blockwise inference of a slightly larger N_e in the Southern population compared to the Northern population. Second, the divergence time inferred by the blockwise analyses corresponds to a period in the PSMC at which the Northern and Southern populations show similar N_e (overlapping confidence intervals), consistent with a shared ancestral population. Finally, PSMC infers an increase in the Northern N_e prior to the time of admixture inferred under the ADM model. Such an increase in genetic diversity in the (Northern) recipient population is exactly what would be expected from a sudden admixture event. We note that PSMC also reveals an increase in the Southern N_e around the same time which, however, is markedly smaller. While PSMC also shows an increase in N_e in the very recent past (i.e. the last 3ky), the variation among bootstrap replicates (Fig. 4) suggests that our data lack the signal to reliably infer population size change over this timescale.

Discussion

Individual level population genomics of very small insects

Genomic data offer enormous potential in inference of population relationships and demography, but genomic resources remain limited for all but a tiny proportion of taxa. In some taxa, including rare species of conservation importance, it is also not possible to sample large panels of individuals (Allendorf et al., 2010; Fuentes-Pardo & Ruzzante

2017). In such cases, and where the distribution of populations is appropriate to a pairwise comparison, the approaches we use here allow demographic inference with minimal cost and sampling. Our approach may be of use in similar pair-wise population analyses of other east coast Australian forest taxa divided by the same dry habitat corridors, and other analogously distributed taxa.

We generated genomic libraries for individual male fig wasps, each only 1.5 mm long, resulting in over 770 thousand aligned blocks of sequence containing an average of 1.5 variable sites. Even with minimal samples of two haploid individuals per population, these whole genome data allowed estimation of demographic parameters with high confidence. A further advantage of the two methods we use is that neither makes any assumptions about phase in sequence data. Both can thus be applied to diploid organisms, provided that coverage is high enough to distinguish sequencing error from heterozygosity (ability to do so with high confidence is a major benefit of working with haploid organisms, including male Hymenoptera).

The best supported IM and ADM models gave similar estimates for the age of the initial divergence between Northern and Southern *P. nigriventris* populations, and their sizes. Further, we show that a burst of admixture (ADM) provides a significantly better fit than ongoing gene flow to the blockwise data and the individual population N_e trajectories inferred by PSMC analysis. We first place our results in the broader context of phylogeographic work on taxa spanning the Burdekin and St. Lawrence Gaps, and then discuss the potential limits of our demographic inferences.

Pleistocene population divergence across the Burdekin and St Lawrence Gaps in *P. nigriventris*

Australia has a complex climate history, a major feature of which is the aridification that started in the Miocene and continued throughout the Pliocene and Pleistocene

(Schauble & Moritz, 2001; Martin, 2006; MacQueen et al., 2010, Frankham et al., 2016).

540 This resulted in the gradual restriction of rain forests to the Eastern coast of Australia,
separated from more arid inland habitats by the Great Dividing Range (Kershaw 1994;
542 McGuigan et al., 1998; Schneider et al., 1998; Pope et al., 2000, Bell et al., 2007).
Between 280 and 205kya, decreased precipitation and more severe aridification were
544 associated with major faunal turnover in rainforest taxa (Hocknull et al., 2007). Against the
backdrop of this general trend, the Pleistocene climate oscillations drove cycles of rain
546 forest expansion during warmer, wetter interglacials and contraction during cooler, drier
glacials (Byrne 2008; Maldonado et al., 2012; Burke et al., 2013). The current dry habitat
548 corridors of the Burdekin and St. Lawrence Gaps are thought to be products of the long
term aridification of Australia, and though it is uncertain when they first formed, they most
550 likely existed through multiple Pleistocene cycles (Bryant & Krosch, 2016).

We found a strong signature of population divergence over the combined Burdekin
552 and St. Lawrence gaps in *Pleistodontes nigriventris*. This is perhaps to be expected, given
the obligate dependence of *P. nigriventris* on *Ficus watkinsiana*, and the restriction of this
554 fig species to rain forest (Dixon 2003). However, our analyses do not allow inference of the
ancestral distribution of *P. nigriventris*, and are compatible with either population being
556 founded from the other, or vicariance of a previously continuous rainforest distribution
(Martin, 2006). Assuming four generations per year for *P. nigriventris* (see below), the best
558 fitting IM and ADM models for *P. nigriventris* both infer divergence between the Northern
and Southern populations 170-200kya ago, in the Late Pleistocene.

560 In a review encompassing a wide range of plant and animal taxa, Bryant and
Krosch (2016) identified a signature of population subdivision for the Burdekin Gap in 18 of
562 27 studies (nine of which have divergence time estimates) and for the St. Lawrence Gap in
10 of 23 studies (six of which have divergence time estimates). Pleistocene divergence
564 across the Burdekin Gap has also been inferred for *Melomys cervinipes* (a wet forest

rodent (Bryant & Fuller (2014)), *Petaurus australis* (yellow-bellied glider (Brown et al.,
566 2006)), and *Varanus varius* (a large lizard with broad habitat preferences (Smitsen et al.,
(2013)). Very few studies have considered insect population structure across the same
568 potential barriers to gene flow. In contrast to our results for *P. nigriventris*, Schiffer et al.,
(2007) found no evidence for genetic divergence across the Burdekin gap in *Drosophila*
570 *birchii*, a specialist rainforest fruit fly, but instead inferred a moderate gene flow across the
whole range following a recent range expansion. Divergence estimates for other taxa
572 spanning the Burdekin and/or St Lawrence Gaps are concentrated in the late Miocene to
late Pleistocene (Bryant & Krosch 2016) with substantial variation across taxa. For
574 example, divergence across the Burdekin Gap was estimated to have occurred in the early
Miocene-late Oligocene >20 million years ago (mya) in *Uperoleia* frogs (Catullo & Keogh
576 2014; Catullo et al., 2014), and 31-51mya in assassin spiders (Rix & Harvey 2012). Given
that most of these previous estimates are not based on any statistical model of population
578 divergence, but rather a single (often mitochondrial) gene tree, it is unclear how much of
the variation in previous divergence time estimates across these co-distributed taxa simply
580 reflects coalescence variance and/or differences in calibration. Model based comparative
studies are needed to assess whether past vicariance events have been shared in time
582 across taxa with disjunct distributions across the Burdekin and St Lawrence gaps.

Two compatible explanations could explain observed population divergence in *P.*
584 *nigriventris*. A first is that aridification and lack of suitable fig hosts in intervening habitats
prevented viable dispersal between Northern and Southern populations. A second is that
586 despite dispersal, migrants failed to contribute genes to receiving populations. This could
happen if, following initial divergence, each of the Northern and Southern populations
588 became locally adapted, but reciprocally maladapted (Rodriguez et al., 2017). Such failure
could result from reciprocal mismatches in climate, or in coevolved interactions with co-
590 distributed *Ficus watkinsiana*. That failure to disperse may not wholly explain divergence

between Northern and Southern populations of *P. nigriventris* is suggested by inferred high
dispersal within each population (low divergence between individuals), and the fact that
some other fig wasps (including another *Pleistodontes* species (Sutton et al. 2016)) show
little or no genetic divergence over distances of around 1000km, close to the separation
between our populations (e.g. Kobmoo et al., 2010; Liu et al., 2015; Tian et al. 2015; Bain
et al., 2016). Further evidence comes from lack of spatial genetic structure (and hence
long range pollen and/or fruit dispersal) in other monoecious figs (e.g. Nazareno et al.,
2013; Bain et al., 2016). That effects other than barriers to dispersal *per se* can structure
intraspecific genetic diversity is also suggested by (potentially host-mediated)
differentiation in other fig wasp species over distances of less than 50km (Tian et al.,
2015). The extent to which Northern and Southern populations of *P. nigriventris* are able to
interbreed and to induce galls in allochthonous host figs remains unknown, but could be
tested experimentally.

Effective population size of *P. nigriventris*

We estimated the effective population sizes for Northern and Southern populations of *P. nigriventris* to be ca. 60k and 70k respectively, with 95% confidence intervals over the IM and ADM models spanning 54k-73k. These figures fall within the range observed in other similarly sized chalcidoid and cynipoid Hymenoptera (e.g. Bunnefeld et al., 2018; Walton et al., 2020). The true census population size for *P. nigriventris* is almost certainly much higher, given that a single large fig tree can bear many thousands of pollinated figs, each of which required entry by at least one female *P. nigriventris*. Census population size (N , which contributes to fruit set) is generally one to several orders of magnitude greater than N_e (Lewontin 1974; Frankham 1995; Palstra and Ruzzante 2008), particularly in taxa that, like fig wasps, show high levels of sib-mating (Kimura and Crow 1963; Leffler et al., 2012; Sutton et al., 2016 for *Pleistodontes*; Molbo et al., 2004 for *Pegoscapus*) and can

experience large population fluctuations and genetic bottlenecks (Bronstein and Hossaert-McKey 1996; Harrison 2000; Wachi et al., 2016). Both are likely explanations for the similar and very low genetic diversities observed in each of the Northern and Southern populations ($\pi = 0.000884$ and 0.000822 respectively), amongst the lowest estimates for any sexually reproducing arthropod (Leffler et al., 2012; Romiguier et al., 2014).

The Northern population of *P. nigriventris* received a burst of admixture from the South at the end of the Pleistocene

Both the best fitting ADM and IM models inferred gene flow from the Southern to the Northern population. While the continuous migration rate inferred under the best IM model ($M = 0.071$) is much lower than the estimated admixture fraction ($f=0.24$) under the better supported ADM model, the overall amount of gene flow inferred under both models is in fact very similar: Under the IM model, the probability that a single lineage sampled from the North is derived from the South via gene flow is $1-e^{-MT}$. Given our estimates for these parameters in model IM9, this is ~ 0.3 , so very comparable to the admixture fraction inferred under the best ADM model. While both models agree in the overall amount of post-divergence gene flow, our model comparison clearly shows that genetic exchange occurred as a sudden burst rather than a continuous process. Additional support for a discrete admixture event comes from the contemporary increase in the Northern population size revealed by PSMC. The inferred combination of divergence and admixture is compatible with the following demographic scenario: (i) Northern and Southern populations were separated 170-200 kya following contraction of suitable habitat and expansion of intervening inhospitable dry forest corridors. (ii) Northern and Southern populations remained separated for over 100ky until favourable conditions in the late Pleistocene allowed expansion of one or both populations of *Ficus watkinsiana*, to the point at which substantial genetic exchange was possible between populations of

pollinating *P. nigriventris*. (iii) Subsequent aridification resulted again in range contractions
644 and a shutdown of gene flow.

Notwithstanding the uncertainty in our time calibration for *P. nigriventris* (see
646 below), the date of the inferred admixture event falls in Marine Isotope Stage 3 (27-60kya),
a period in the late Pleistocene characterised globally by abrupt phases of warming and
648 cooling (Siddall et al., 2008; Van Meerbeeck et al., 2009). These warming phases were
interspersed by cooler periods around every 7,000 years (Clark et al., 2007), and even
650 during the cooler periods average temperatures were much higher than during the Last
Glacial Maximum (~19-21kya) (Van Meerbeeck et al., 2009). The lower bound (53kya) of
652 the inferred admixture time corresponds approximately to the warmest point in one of
these cycles, and it is tempting to suggest that this allowed temporary expansion of the
654 range of the *F. watkinsiana*/*P. nigriventris* mutualism and secondary genetic contact. Such
climatic instability also raises the possibility that any selection in favour of locally adapted
656 fig wasps could sometimes have been replaced by selection in favour of adaptive
introgression of migrant genes (Hedrick 2013).

658 Our results do not rule out bidirectional dispersal, but rather imply a signal of
predominant gene flow from the South into the North. This could indicate conditions that
660 facilitated northwards dispersal in this direction, such as prevailing winds from the south.
Records from the 1940s-2000s do indicate stronger winds from the south in eastern
662 Australia (Australian Government, Bureau of Meteorology,
<http://www.bom.gov.au/climate/>). However, it is not known whether current trends can be
664 extended back into the past. Similar post-divergence dispersal from south to north across
the Burdekin gap has been inferred in a small number of other rain forest-associated taxa
666 (Bryant & Fuller, 2014; Bryant & Krosch 2016). Alternatively, the asymmetry in genetic
exchange between North and South we infer may be the result of genetic incompatibilities
668 that arose and became fixed during periods of isolation.

670 **Limits of demographic inference**

Our inference of the population history of *P. nigriventris* is contingent on the realism
672 of our models, ability to incorporate effects of linkage, and scaling of time estimates. We
consider these issues in turn.

674 Our IM and ADM models assume two populations with a maximum of two different
and constant N_e parameters (Fig. 1). While this simple model is unlikely to be true for any
676 organism, the assumption of two panmictic populations is, as far as is known, a good
approximation for the current distribution of *P. nigriventris*. Fitting explicit models
678 necessarily involves simplifying assumptions, and alternatives for limiting the number of N_e
parameters (such as assuming equal N_e for all populations, or an even split between
680 daughter populations) are even less realistic. Concordance in N_e estimates between our
IM/ADM models and our PSMC analysis (in which N_e is free to vary through time)
682 suggests that drastic changes in N_e are not a feature of the population history of *P.*
nigriventris, and that this simplifying assumption is unlikely to affect our inference.
684 Agreement in parameter estimates across models and the close fit of our best ADM model
to the observed blockwise data both suggest that we have captured key aspects of the
686 demographic history of *P. nigriventris*. Perhaps the main result of our analyses is that we
have shown that continuous and discrete gene flow can be clearly distinguished in this
688 species, even over relatively recent timescales (both in terms of sequence divergence and
genetic drift). While most analyses of demographic history choose *a priori* to model gene
690 flow as either a continuous process or a discrete event, these two extremes of model
space are rarely compared directly and it remains an open question which one better
692 captures the demographic history in most taxa. All three demographic scenarios we have
considered (IM, ADM and PSMC) are nevertheless crude simplifications of what is likely a
694 more complex history, possibly involving repeated cycles of rain forest expansion and

contraction. The upland wet forests south of the Burdekin Gap in the Clarke Range and
696 Conway Peninsula are thought to have persisted through multiple Pleistocene climate
cycles (Stuart-Fox et al., 2001), and it would be interesting in future to use the information
698 contained in larger samples to explore more realistic histories involving repeated
admixture pulses (Jésus et al., 2006) in *P. nigriventris* and *F. watkinsiana*, potentially
700 involving additional intermediate populations that may no longer exist (e.g. Stone et al.,
2017). Likewise, it would be interesting to test how well the blockwise distributions of
702 divergence and diversity in *P. nigriventris* fit a more complex generalised class of IM
models (GIM) that involve periods of historic gene flow (Costa & Wilkinson-Herbots 2017).
704 Exploring these models would allow bridging of the model space between the IM and ADM
histories we consider here (Costa & Wilkinson-Herbots 2017). However, unlike the bSFS
706 framework we have used here, analytic results for GIM models are currently limited to
pairwise samples, which, all else being equal, are less informative about gene flow (Lohse
708 et al., 2016, Fig. 7).

Demographic analyses that are based on genome-wide samples of either single
710 variants or loci/blocks generally assume that blocks are statistically independent, i.e.
unlinked. A common way of dealing with linkage effects is to use subsampling, either by
712 sampling a fixed minimum distance apart or by resampling bootstrap. We have instead
opted for a fully parametric bootstrap that incorporates recombination within and linkage
714 between blocks explicitly. Although this is computationally more intensive than
subsampling, we believe it is the only way to accurately capture the effect on parameter
716 estimates. Our parametric bootstrap for the best fitting IM and ADM models gave narrow
95% confidence intervals for all model parameters. Importantly, mean parameter estimates
718 obtained from the simulation replicates are very close to the MLEs used to simulate the
datasets (Table 1). This confirms that biases in parameter estimates due to violations of
720 the assumption of no recombination within blocks are negligible, as shown in previous

studies on a range of organisms (Jennings & Edwards, 2005; Lanier & Knowles, 2012; Hearn et al., 2014; Bunnefeld et al., 2015; Wang & Liu, 2016). Another standard assumption of demographic inference is that sequences evolve neutrally. While our approach sampled blockwise sequence variation genome-wide, one would expect both genome assembly and read mapping to be easier in regions under selective constraint. As a consequence, our blockwise dataset is likely enriched for conserved coding regions. In a methodologically similar study of a European gall wasp, Hearn et al., (2014) tested for the effect of selective constraint by partitioning blocks according to the proportion of coding sequence they contained, and scaling the estimated genome-wide mutation rate between values for synonymous and non-synonymous mutations. They found that mutation rate heterogeneity had no impact on the inference, reporting only a slight increase in N_e and divergence time estimates. Given these results and the lack of annotated genomes or transcriptome data to aid gene detection, we have not pursued this here. However, it will be interesting to check whether selective constraint can explain the observed higher frequency of invariant blocks in our *P. nigriventris* data compared to expectations under both the best fitting IM and ADM models. Taken together, these results suggest that recombination within blocks and background selection had a minimal impact on our inference of demographic history.

Scaling divergence and admixture times into years requires knowledge of both the mutation rate and generation time. Both are uncertain for *P. nigriventris*. In the absence of direct mutation rate estimates for fig wasps, we used an estimate from *Drosophila melanogaster* (Keightley et al., 2014). While it is unclear how well this matches mutation rates in fig wasps, it is reassuring that the few published mutation rate estimates for other insects are similar (e.g. 2.9×10^{-9} for the butterfly *Heliconius melpomene*). A likely greater source of calibration uncertainty is average wasp generation time, which is not known for *P. nigriventris*. Our estimate of 4 generations per year is based on data for other Australian

Pleistodontes species, taking into account the seasonal habitat and large (slowly
748 developing) fruits of *F. watkinsiana* (see Methods). Calibration of the timing of
demographic events in *P. nigriventris* scales simply and inversely with generation time,
750 i.e., halving the number of generations per year doubles the age of events. Assuming 2
and 6 generations per year (rather than 4 as we have done here) still places divergence
752 and admixture times for *P. nigriventris* in the late Pleistocene: for 2 generations/year,
 $T=392(386-396)$ ky and $T_{adm}=114(106-184)$ ky, and for 6 generations/year, $T=131(129-$
754 $132)$ ky and $T_{adm}= 38(35-61)$ ky. However, in the absence of precise information our attempt
to match the demographic history of *P. nigriventris* to past climatic events remains
756 speculative.

758 Our results show that robust demographic inferences can be made for very small
sample sizes of non-model taxa without significant associated genomic resources. While
760 our motivation for assembling a genome for *P. nigriventris* was solely to generate a
dataset for demographic inference, genome assemblies are arguably a more generally
762 useful resource than other types of genomic data (e.g. RadSeq) and can be improved in
the future (e.g. by adding RNASeq and long read data to improve annotation and
764 contiguity respectively). Inference based on de novo assemblies and minimal sampling is
likely to become the norm for taxa that are rare or hard to sample. Figs support highly
766 structured (and often co-evolved) communities of pollinators, inquilines and natural
enemies (Lopez-Vaamonde et al., 2001; Segar et al., 2014), and while a growing body of
768 work addresses phylogeographic relationships in figs and their pollinators (Bain et al.,
2016; Rodriguez et al., 2017; Yu et al., 2019), much less is known about the
770 phylogeography of non-pollinating fig wasps (Sutton et al., 2016). Our approach provides a
framework for comparative analyses that reconstruct the assembly of these species-rich
772 communities. More broadly, given that many questions about both intraspecific

demography history and speciation come down to distinguishing between ongoing gene
774 flow and discrete admixture pulses, systematic power analyses on how this can best be
achieved - especially for genomic data from non-model organisms without a contiguous
776 reference genome - are urgently needed.

778 **Acknowledgements**

This work was funded by a UK NERC grant (NE/J010499) to GS, KL and JMC. LC was
780 supported by a PhD studentship from the UK NERC. KL was supported by a NERC
fellowship (NE/L011522/1)

782 **References**

- 784 Aeschbacher, S., Selby, J. P., Willis, J. H. & Coop, G. (2017). Population-genomic
inference of the strength and timing of selection against gene flow. *Proceedings of the
National Academy of Sciences of the USA*, 114, 7061-7066.
786 <https://doi.org/10.1073/pnas.1616755114>
- 788 Ahmed, S., Compton, S. G., Butlin, R. K., & Gilmartin P. M. (2009). Wind-borne insects
mediate directional pollen transfer between desert fig trees 160 kilometers apart.
790 *Proceedings of the National Academy of Sciences of the USA*, 106, 20342-20347.
<https://doi.org/10.1073/pnas.0902213106>
- 792 Al-beidh, S. (2010). Investigations into stability in the fig/fig-wasp mutualism. Unpublished
794 PhD thesis, Imperial College, London, U.K.
- 796 Allendorf, F. W., Hohenlohe, P. A & Luikart, G. (2010). Genomics and the future of
conservation genetics. *Nature Reviews Genetics*, 11, 697–709.
798 <https://doi.org/10.1038/nrg2844>
- 800 Bain, A., Borges, R. M., Chevallier, M. H., Vignes, H., Kobmoo, N., Peng, Y. Q., Cruaud,
A., Rasplus, J. Y., Kjellberg, F., & Hossaert-Mckey, M. (2016). Geographic structuring into
802 vicariant species-pairs in a wide-ranging, high-dispersal plant–insect mutualism: the case
of *Ficus racemosa* and its pollinating wasps. *Evolutionary Ecology*, 30, 663–684.
804 <https://doi.org/10.1007/s10682-016-9836-5>
- 806 Bankevich, A., Nurk, S., Antipov, D., Gurevich, A. A., Dvorkin, M., Kulikov, A. S. &
Pevzner, P. A. (2012). SPAdes: a new genome assembly algorithm and its applications to
808 single cell sequencing. *Journal of Computational Biology*, 19, 455-477.
<https://doi.org/10.1089/cmb.2012.0021>
- 810 Baker, C. H., Graham, G. C., Scott, K. D., Cameron, S. L., Yeates, D. K. & Merritt, D. J.
812 (2008). Distribution and phylogenetic relationships of Australian glow-worms

814 *Arachnocampa* (Diptera, Keroplatidae)', *Molecular Phylogenetics and Evolution*, 48, 506-
514. <https://doi.org/10.1016/j.ympev.2008.04.037>

816 Bell, K. L., Moritz, C., Moussalli, A. & Yeates, D. K. (2007). Comparative phylogeography
and speciation of dung beetles from the Australian Wet Tropics rainforest. *Molecular*
818 *Ecology*, 16, 4984-4998. <https://doi.org/10.1111/j.1365-294X.2007.03533.x>

820 The impact of global selection on local adaptation and reproductive isolation
Bisschop, G., Setter, D., Rafajlović, M., Baird, S. J. E. & Lohse, K. (2019). bioRxiv 855320.
822 <https://doi.org/10.1101/855320>

824 Bronstein, J. L., Gouyon, P.-H., Gliddon, C., Kjellberg, F. & Michaloud, G. (1990).
The ecological consequences of flowering asynchrony in monoecious figs: a simulation
826 study. *Ecology*, 71, 2145-2156. <https://doi.org/10.2307/1938628>

828 Bronstein, J. L. & Hossaert-McKey, M. (1996). Variation in reproductive success within a
subtropical fig/pollinator mutualism. *Journal of Biogeography*, 23, 433–446.
830 <https://doi.org/10.1111/j.1365-2699.1996.tb00005.x>

832 Brown, M., Cooksley, H., Carthew, S. M. & Cooper, S. J. B. (2006). Conservation units and
phylogeographic structure of an arboreal marsupial, the yellow-bellied glider (*Petaurus*
834 *australis*). *Australian Journal of Zoology*, 54, 305-317. <https://doi.org/10.1071/ZO06034>

836 Bryant, L. M. & Fuller, S. J. (2014). Pleistocene climate fluctuations influence
phylogeographical patterns in *Melomys cervinipes* across the mesic forests of eastern
838 Australia. *Journal of Biogeography*, 41, 1923-1935. <https://doi.org/10.1111/jbi.12341>

840 Bryant, L. M. & Krosch, M. N. (2016). Lines in the land: a review of evidence for eastern
Australia's major biogeographical barriers to closed forest taxa. *Biological Journal of the*
842 *Linnean Society*, 119, 238-264. <https://doi.org/10.1111/bij.12821>

844 Buchfink, B., Xie, C. & Huson, D. H. (2015). Fast and sensitive protein alignment using
DIAMOND. *Nature Methods*, 12, 59-60. <https://doi.org/10.1038/nmeth.3176>
846

Bunnefeld, L., Hearn, J., Stone, G.N. & Lohse, K. (2018). Whole genome data reveal the
848 complex history of a diverse ecological community. *Proceedings of the National Academy*
of Sciences of the United States of America. 115, E6507-E6515.
850 <https://doi.org/10.1073/pnas.1800334115>

852 Burke, J. M., Ladiges, P.Y., Batty, E. L., Adams, P. B., Bayly, M. J. & Katinas, L. (2013).
Divergent lineages in two species of *Dendrobium* orchids (*D. speciosum* and *D.*
854 *tetragonum*) correspond to major geographical breaks in eastern Australia. *Journal of*
Biogeography, 40, 2071-2081. <https://doi.org/10.1111/jbi.12145>
856

Byrne, M. (2008). Evidence for multiple refugia at different time scales during Pleistocene
858 climatic oscillations in southern Australia inferred from phylogeography. *Quaternary*
Science Reviews, 27, 2576-2585.
860

Catullo, R. A. & Keogh, J. S. (2014). Aridification drove repeated episodes of
862 diversification between Australian biomes: evidence from a multi-locus phylogeny of
Australian toadlets (*Uperoleia*: Myobatrachidae). *Molecular Phylogenetics and Evolution*,
864 79, 106–117. <https://doi.org/10.1016/j.ympev.2014.06.012>

- 866 Catullo, R. A., Lanfear, R., Doughty, P. & Keogh, J. S. (2014). The biogeographical
868 boundaries of northern Australia: evidence from ecological niche models and a multi-locus
870 phylogeny of *Uperoleia* toadlets (Anura: Myobatrachidae). *Journal of Biogeography*, 41,
659–672. <https://doi.org/10.1111/jbi.12230>
- 872 Chapple, D. G., Hoskin, C. J., Chapple, S. N. & Thompson, M. B. (2011). Phylogeographic
874 divergence in the widespread delicate skink (*Lampropholis delicata*) corresponds to dry
habitat barriers in eastern Australia. *BMC Evolutionary Biology*, 11, 191.
<https://doi.org/10.1186/1471-2148-11-191>
- 876 Chen, Y., Compton, S. G., Liu, M. & Chen, X. Y. (2012). Fig trees at the northern limit of
878 their range: the distributions of cryptic pollinators indicate multiple glacial refugia.
Molecular Ecology, 21, 1687–1701. <https://doi.org/10.1111/j.1365-294X.2012.05491.x>
- 880 Clark, P. U., Hostetler, S. W., Pisias, N. G., Schmittner A. & Meissner, K. J. (2007).
882 Mechanisms for an ~7-Kyr climate and sea-level oscillation during marine isotope stage 3.
In: *Ocean Circulation: Mechanisms and Impacts—Past and Future Changes of Meridional
884 Overturning*. American Geophysical Union. *Geophysical Monograph Series*, Volume 173.
<https://doi.org/10.1029/173GM15>
- 886 Cook, James M. & Rasplus, J.-Y. (2003). Mutualists with attitude: coevolving fig wasps
888 and figs. *Trends in Ecology & Evolution*, 18, 241–248. [https://doi.org/10.1016/S0169-5347\(03\)00062-4](https://doi.org/10.1016/S0169-5347(03)00062-4)
- 890 Cooper, L., Bunnefeld, L., Hearn, J., Cook, J. M., Lohse, K. & Stone, G. N. (2020)a.
892 Phylogeography of *Pleistodontes nigriventris* from northern and southern populations in
Australia. Sequence read data and *SPAdes* genome assembly in the European Nucleotide
894 Archive, Study PRJEB35527. Available from
<https://www.ebi.ac.uk/ena/browser/view/PRJEB35527>
- 896 Cooper, L., Bunnefeld, L., Hearn, J., Cook, J. M., Lohse, K. & Stone, G. N. (2020)b.
898 Phylogeography of *Pleistodontes nigriventris* from northern and southern populations in
Australia. Trimmed sequence block data and associated *Mathematica* notebook for
900 blockwise analyses. Available from the Dryad Digital Repository, doi:// (to be added on
acceptance).
- 902 Costa, R. J. & Wilkinson-Herbots, H. (2017). Inference of gene flow in the process of
904 speciation: an efficient maximum-likelihood method for the isolation-with-initial-migration
model. *Genetics*, 205, 1597–1618. <https://doi.org/10.1534/genetics.116.188060>.
- 906 Darwell, C. T., Al-Beidh, S. & Cook, J. M. (2014). Molecular species delimitation of a
908 symbiotic fig-pollinating wasp species complex reveals extreme deviation from reciprocal
partner specificity. *BMC Evolutionary Biology*, 14, 189. <https://doi.org/10.1186/s12862-014-0189-9>
- 910 DePristo, M. A., Banks, E., Poplin, R., Garimella, K. V., Maguire, J. R., Hartl, C. ... & Daly,
912 M. J. (2011). A framework for variation discovery and genotyping using next-generation
DNA sequencing data. *Nature Genetics*, 43, 491–498. <https://doi.org/10.1038/ng.806>
- 914

- Dixon, D. J. (2003). A taxonomic revision of the Australian *Ficus* species in the section Malvanthera (*Ficus* subg. *Urostigma*: Moraceae). *Telopea*, 10, 125-153. <https://doi.org/10.7751/telopea20035611>
- Dolman, G. & Moritz, C. (2006). A multilocus perspective on refugial isolation and divergence in rainforest skinks (*Carlia*). *Evolution*, 60, 573-582. <https://doi.org/10.1554/05-487.1>
- Durand E. Y., Patterson N., Reich D., & Slatkin, M. (2011). Testing for ancient admixture between closely related populations. *Molecular Biology and Evolution*, 28, 2239–2252. . <https://doi.org/10.1093/molbev/msr048>
- Felsenstein, J. 2003. *Inferring Phylogenies*. Sinauer, 664pp.
- Frankham, R. (1995). Effective population size/adult population size ratios in wildlife: a review. *Genetics Research*, 66, 95–107. <https://doi.org/10.1017/S0016672300034455>
- Frankham, G. J., Handasyde, K. A. & Eldridge, M. D. B (2016). Evolutionary and contemporary responses to habitat fragmentation detected in a mesic zone marsupial, the long-nosed potoroo (*Potorous tridactylus*) in south-eastern Australia. *Journal of Biogeography*, 43, 653-665. <https://doi.org/10.1111/jbi.12659>
- Fuentes-Pardo, A. P. & Ruzzante, D. E. (2017). Whole-genome sequencing approaches for conservation biology: Advantages, limitations and practical recommendations. *Molecular Ecology*, 26, 5369–5406. <https://doi.org/10.1111/mec.14264>
- García-Alcalde, F., Okonechnikov, K., Carbonell, J., Cruz, L. M., Götz, S., Tarazona, S., Dopazo, J., Meyer, T. F. & Conesa, A. (2012). Qualimap: evaluating next-generation sequencing alignment data. *Bioinformatics*, 28, 2678–2679. <https://doi.org/10.1093/bioinformatics/bts503>
- Greeff, J. M., van Vuuren, G. J. J., Kryger, P. & Moore, J. C. (2009). Outbreeding and possibly inbreeding depression in a pollinating fig wasp with a mixed mating system. *Heredity*, 102, 349-356. <https://doi.org/10.1038/hdy.2009.2>
- Haine, E. R., & Cook, J. M. (2005). Convergent incidences of *Wolbachia* infection in fig wasp communities from two continents. *Proceedings of the Royal Society of London, Series B Biological Sciences*, 272, 421-429. <https://doi.org/10.1098/rspb.2004.2956>
- Hare, M. P. (2001). Prospects for nuclear gene phylogeography. *Trends in Ecology & Evolution*, 16, 700-706. [https://doi.org/10.1016/S0169-5347\(01\)02326-6](https://doi.org/10.1016/S0169-5347(01)02326-6)
- Harrison, R. D. (2000). Repercussions of El Niño: drought causes extinction and the breakdown of mutualism in Borneo. *Proceedings of the Royal Society of London, Series B Biological Sciences*, 267, 911–915. <https://doi.org/10.1098/rspb.2000.1089>
- Harrison, R. D. (2003). Fig wasp dispersal and the stability of a keystone plant resource in Borneo. *Proceedings of the Royal Society of London, Series B Biological Sciences*, 270, Suppl 1: S76-9.
- Harrison, R. D. (2005). Figs and the diversity of tropical rainforests. *Bioscience*, 55, 1053-1064. <https://doi.org/10.1098/rsbl.2003.0018>

- 968 Harrison, R. D. & Rasplus, J.-Y. (2006). Dispersal of fig pollinators in Asian tropical rain
forests. *Journal of Tropical Ecology*, 22, 631-639.
970 <https://doi.org/10.1017/S0266467406003488>
- 972 Hearn, J., Stone, G. N., McInnes, L., Nicholls, J. A., Barton, N. H. & Lohse, K. (2014).
Likelihood-based inference of population history from low coverage de novo genome
974 assemblies. *Molecular Ecology*, 23, 198-211. <https://doi.org/10.1111/mec.12578>.
- 976 Hedrick, P. W. (2013). Adaptive introgression in animals: examples and comparison to
new mutation and standing variation as sources of adaptive variation. *Molecular Ecology*,
978 22, 4606-4618. <https://doi.org/10.1111/mec.12415>
- 980 Hey, J. & Nielsen, R. (2004). Multilocus methods for estimating population sizes, migration
rates and divergence time, with applications to the divergence of *Drosophila*
982 *pseudoobscura* and *D. persimilis*. *Genetics*, 167, 747-60.
<https://doi.org/10.1534/genetics.103.024182>.
- 984 Hocknull, S. A., Zhao, J.-X., Feng, Y.-X. & Webb, G. E. (2007). Responses of Quaternary
rainforest vertebrates to climate change in Australia. *Earth and Planetary Science Letters*,
986 264, 317-331. <https://doi.org/10.1016/j.epsl.2007.10.004>
- 988 Jennings, W. B. & Edwards, S. V. (2005). Speciation history of Australian grass finches
(*Poephila*) inferred from thirty gene trees. *Evolution*, 59, 2033-2047.
990 <https://doi.org/10.1111/j.0014-3820.2005.tb01072.x>
- 992 Jermini, L. S. & Crozier, R. H. (1994). The cytochrome b region in the mitochondrial DNA
of the ant *Tetraponera rufoniger*: sequence divergence in Hymenoptera may be associated
994 with nucleotide content. *Journal of Molecular Evolution*, 38, 282-294.
996 <https://doi.org/10.1007/BF00176090>
- 998 Jesus, F. F., Wilkins, J. F., Solferini, V. N. & Wakeley, J. (2006). Expected coalescence
times and segregating sites in a model of glacial cycles. *Genetic and Molecular Research*,
1000 5, 466-474.
- 1002 Jia, X. C., Yao, J. Y., Chen, Y. Z., Cook, J. M & Crozier, R. H. (2008). The phenology and
potential for self-pollination of two Australian monoecious fig species. *Symbiosis*, 45, 91–
1004 96.
- 1006 Keightley, P. D., Ness, R. W., Halligan, D. L. & Haddrill, P. R. (2014). Estimation of the
spontaneous mutation rate per nucleotide site in a *Drosophila melanogaster* full-sib family.
1008 *Genetics*, 196, 313-320. <https://doi.org/10.1534/genetics.113.158758>
- 1010 Keightley, P. D., Pinharanda, A., Ness, R. W., Simpson, F., Dasmahapatra, K. K., Mallet,
J., Davey, J. W. & Jiggins C. D. (2015). Estimation of the spontaneous mutation rate in
1012 *Heliconius melpomene*. *Molecular Biology and Evolution*, 32, 239-243.
<https://doi.org/10.1093/molbev/msu302>
- 1014 Kelleher, J., Etheridge, A. M. & McVean, G. (2016). Efficient coalescent simulation and
genealogical analysis for large sample sizes. *PLoS Computational Biology*, 12, e1004842.
1016 <https://doi.org/10.1371/journal.pcbi.1004842> <https://doi.org/10.1371/journal.pcbi.1004842>
- 1018

- Kershaw, A. P. (1994). Pleistocene vegetation of the humid tropics of northeastern Queensland, Australia. *Paleoclimatology, Paleogeography, Paleoecology*, 109, 399–412. [https://doi.org/10.1016/0031-0182\(94\)90188-0](https://doi.org/10.1016/0031-0182(94)90188-0) [https://doi.org/10.1016/0031-0182\(94\)90188-0](https://doi.org/10.1016/0031-0182(94)90188-0)
- Kimura, M. & Crow, J. F. (1963). The measurement of effective population number. *Evolution*, 17, 279-288. <https://doi.org/10.1111/j.1558-5646.1963.tb03281.x>
- Kobmoo, N., Hossaert-McKey, M., Rasplus, J.-Y. & Kjellberg, F. (2010). *Ficus racemosa* is pollinated by a single population of a single agaonid wasp species in continental South-East Asia. *Molecular Ecology*, 19, 2700-2712. <https://doi.org/10.1111/j.1365-294X.2010.04654.x>
- Kumar, S., Jones, M., Koutsovoulos, G., Clarke, M. & Blaxter, M. (2013). Blobology: exploring raw genome data for contaminants, symbionts and parasites using taxon-annotated GC-coverage plots', *Frontiers in Genetics*, 4, 237. <https://doi.org/10.3389/fgene.2013.00237>
- Langmead, B., & Salzberg, S. L. (2012). Fast gapped-read alignment with Bowtie 2. *Nature Methods*, 9, 357-359. <https://doi.org/10.1038/nmeth.1923>
- Lanier, H. C., & Knowles, L. L.(2012). Is recombination a problem for species-tree analyses? *Systematic Biology*, 61, 691-701. <https://doi.org/10.1093/sysbio/syr128>
- Leffler, E. M., Bullaughey, K., Matute, D. R., Meyer, W. K., Ségurel, L., Venkat, A., Andolfatto, P. & Przeworski, M. (2012). Revisiting an old riddle: what determines genetic diversity levels within species? *PLoS Biology*, 10(9), e1001388. <https://doi.org/10.1371/journal.pbio.1001388>
- Lewontin, R. C. (1974). *The Genetic Basis of Evolutionary Change*. Columbia University Press, New York. 346 pp.
- Li, H. (2011). A statistical framework for SNP calling, mutation discovery, association mapping and population genetical parameter estimation from sequencing data. *Bioinformatics*, 27, 2987-2993. <https://doi.org/10.1093/bioinformatics/btr509>
- Li, H. & Durbin, R. (2009). Fast and accurate short read alignment with Burrows-Wheeler transform. *Bioinformatics*, 25, 1754-1760. <https://doi.org/10.1093/bioinformatics/btp324>
- Li, H. & Durbin, R. (2011). Inference of human population history from individual whole-genome sequences. *Nature*, 475, 493-496. <https://doi.org/10.1038/nature10231>
- Li, H., Handsaker, B., Wysoker, A., Fennell, T., Ruan, J., Homer, N., Marth, G., Abecasis, G., Durbin, R. & Subgroup Genome Project Data Processing. (2009). The Sequence Alignment/Map format and SAMtools. *Bioinformatics*, 25, 2078-2079. <https://doi.org/10.1093/bioinformatics/btp352>
- Liu, M., Zhao, R., Chen, Y., Zhang, J., Compton, S. G. & Chen, X. Y. (2014). Competitive exclusion among fig wasps achieved via entrainment of host plant flowering phenology. *PLoS ONE* 9(5): e97783. doi:10.1371/journal.pone.0097783
- Liu, M., Compton, S. G., Peng, F. E., Zhang, J. & Chen, X. Y. (2015). Movements of genes between populations: are pollinators more effective at transferring their own or plant

- genetic markers?', *Proceedings of the Royal Society of London, Series B Biological Sciences*, 282, 20150290. <https://doi.org/10.1098/rspb.2015.0290>
- Lohse, K., Chmelik, M., Martin, S. H. & Barton, N. H. (2016). Efficient strategies for calculating blockwise likelihoods under the coalescent. *Genetics*, 202, 775-786. <https://doi.org/10.1534/genetics.115.183814>
- Lohse, K. & Frantz, L. A. F. (2014). Neandertal admixture in Eurasia confirmed by maximum-likelihood analysis of three genomes. *Genetics*, 196, 1241-1251. <https://doi.org/10.1534/genetics.114.162396>
- Lohse, K., Harrison, R. J. & Barton, N. H. (2011). A general method for calculating likelihoods under the coalescent process. *Genetics*, 189, 977-987. <https://doi.org/10.1534/genetics.111.129569>
- Lopez-Vaamonde, C., Rasplus, J.Y., Weiblen, G.D., and Cook, J.M. (2001) Molecular phylogenies of fig wasps: Partial cocladogenesis of pollinators and parasites. *Molecular Phylogenetics and Evolution*, 21, 55-71. <https://doi.org/10.1006/mpev.2001.0993>
- Lopez-Vaamonde, C., Dixon, D. J., Cook, J. M. & Rasplus, J.-Y. (2002). Revision of the Australian species of *Pleistodontes* (Hymenoptera: Agaonidae) fig-pollinating wasps and their host-plant associations. *Zoological Journal of the Linnean Society*, 136, 637-683. <https://doi.org/10.1046/j.1096-3642.2002.00040.x>
- Macqueen, P., Seddon, J. M., Austin, J. J., Hamilton, S. & Goldizen, A. W. (2010). Phylogenetics of the pademelons (Macropodidae: *Thylogale*) and historical biogeography of the Australo-Papuan region. *Molecular Phylogenetics and Evolution*, 57, 1134-1148. <https://doi.org/10.1016/j.ympev.2010.08.010>
- Macqueen, P., Seddon, J. M. & Goldizen, A. W. (2012). Effects of historical forest contraction on the phylogeographic structure of Australo-Papuan populations of the red-legged pademelon (Macropodidae: *Thylogale stigmatica*). *Austral Ecology*, 37, 479-490. <https://doi.org/10.1111/j.1442-9993.2011.02309.x>
- Maldonado, S. P., Melville, J., Peterson, G. N. L. & Sumner, J. (2012). Human-induced versus historical habitat shifts: identifying the processes that shaped the genetic structure of the threatened grassland legless lizard, *Delma impar*. *Conservation Genetics*, 13, 1329-1342. <https://doi.org/10.1007/s10592-012-0377-3>
- Male, T. D. & Roberts, G. E. (2005). Host associations of the strangler fig *Ficus watkinsiana* in a subtropical Queensland rain forest. *Austral Ecology*, 30, 229-236. <https://doi.org/10.1111/j.1442-9993.2005.01442.x>
- Markgraf, V., McGlone, M. & Hope, G. (1995). Neogene paleoenvironmental and paleoclimatic change in southern temperate ecosystems — a southern perspective. *Trends in Ecology & Evolution*, 10, 143-147. [https://doi.org/10.1016/S0169-5347\(00\)89023-0](https://doi.org/10.1016/S0169-5347(00)89023-0)
- Martin, H. A. (2006). Cenozoic climatic change and the development of the arid vegetation in Australia. *Journal of Arid Environments*, 66, 533-563. <https://doi.org/10.1016/j.jaridenv.2006.01.009>

1124 Martin, M. (2011). Cutadapt removes adapter sequences from high-throughput sequencing
reads. *EMBnet.journal*, 17, No 1: Next Generation Sequencing Data Analysis.
https://doi.org/https://doi.org/10.14806/ej.17.1.200

1126 Martin, S. H., Dasmahapatra, K. K., Nadeau, N. J., Salazar, C., Walters, J. R., Simpson,
1128 F., Blaxter, M., Manica, A., Mallet, J. & Jiggins, C. D. (2013). Genome-wide evidence for
speciation with gene flow in *Heliconius* butterflies. *Genome Research*, 23, 1817-1828.
1130 https://doi.org/10.1101/gr.159426.113

1132 McGuigan, K., McDonald, K., Parris, K. & Moritz, C. (1998). Mitochondrial DNA diversity
and historical biogeography of a wet forest-restricted frog (*Litoria pearsoniana*) from mid-
1134 east Australia. *Molecular Ecology*, 7, 175-186. https://doi.org/10.1046/j.1365-
294x.1998.00329.x

1136 McKenna, A., Hanna, M., Banks, E., Sivachenko, A., Cibulskis, K., Kernytsky, A.,
1138 Garimella, K., Altshuler, D., Gabriel, S., Daly, M. & DePristo, M. A. (2010). The Genome
Analysis Toolkit: a MapReduce framework for analyzing next generation DNA sequencing
1140 data. *Genome Research*, 20, 1297-1303. https://doi.org/10.1101/gr.107524.110

1142 Molbo, D., Machado, C. A., Herre, E. A. & Keller, L. (2004). Inbreeding and population
structure in two pairs of cryptic fig wasp species. *Molecular Ecology*, 13, 1613–1623.
1144 https://doi.org/10.1111/j.1365-294X.2004.02158.x

1146 Nazareno, A. G., Alzate-Marin, A. L. & Pereira, R. A. S. (2013). Dioecy, more than
monoecy, affects plant spatial genetic structure: the case study of *Ficus*. *Ecology and*
1148 *Evolution*, 3, 3495–3508. https://doi.org/doi: 10.1002/ece3.739

1150 Nicholls, J. A., & Austin, J. J. (2005). Phylogeography of an east Australian wet forest bird,
the satin bowerbird (*Ptilonorhynchus violaceus*), derived from mtDNA, and its relationship
1152 to morphology. *Molecular Ecology*, 14, 1485-1496. https://doi.org/10.1111/j.1365-
294X.2005.02544.x

1154 Nielsen, R. & Wakeley, J. (2001). Distinguishing migration from isolation: a Markov Chain
Monte Carlo approach. *Genetics*, 158, 885-896.

1156 Niu, L.-H., Song, X.-F., He, S. M., Zhang, P., Wang, N. X., Li, Y. & Huang, D. W. (2015).
1158 New insights into the fungal community from the raw genomic sequence data of fig wasp
Ceratosolen solmsi. *BMC Microbiology*, 15, 27. https://doi.org/10.1186/s12866-015-0370-3

1160 Nürnberger, B., Lohse, K., Fijarczyk, A., Szymura, J. M. & Blaxter, M. L. (2016). Para-
1162 allopatry in hybridizing fire-bellied toads (*Bombina bombina* and *B. variegata*): inference
from transcriptome-wide coalescence analyses. *Evolution*, 70, 1803-1818.
1164 https://doi.org/10.1111/evo.12978

1166 Okonechnikov, K., Conesa, A. & García-Alcalde, F (2016). Qualimap 2: advanced multi-
sample quality control for high-throughput sequencing data. *Bioinformatics*, 32, 292–294.
1168 https://doi.org/10.1093/bioinformatics/btv566

1170 Oswald, J. A., Overcast, I., Mauck, W. M., Andersen, M. J. & Smith, B. T. (2017). Isolation
with asymmetric gene flow during the nonsynchronous divergence of dry forest birds.
1172 *Molecular Ecology*, 26, 1386-1400. https://doi.org/10.1111/mec.14013

- 1174 Palstra, F. P. & Ruzzante, D. E. (2008). Genetic estimates of contemporary effective
1176 population size: what can they tell us about the importance of genetic stochasticity for wild
population persistence? *Molecular Ecology*, 17, 3428-3447.
<https://doi.org/10.1111/j.1365-294X.2008.03842.x>
- 1178 Pope, L. C., Estoup, A. & Moritz, C. (2000). Phylogeography and population structure of an
1180 ecotonal marsupial, *Bettongia tropica*, determined using mtDNA and microsatellites.
Molecular Ecology, 9, 2041-2053. <https://doi.org/10.1046/j.1365-294x.2000.01110.x>
- 1182 Pope, L. C., Storch, D., Adams, M., Moritz, C. & Gordon, G. (2001). A phylogeny for the
1184 genus *Isoodon* and a range expansion for *I. obesulus peninsulae* based on mtDNA control
region and morphology. *Australian Journal of Zoology*, 49, 411-434.
1186 <https://doi.org/10.1071/ZO00060>
- 1188 Price, A. L., Jones, N. C. & Pevzner, P. A. (2005). *De novo* identification of repeat families
in large genomes. *Bioinformatics*, 21 Suppl 1, i351-i358.
1190 <https://doi.org/10.1093/bioinformatics/bti1018>
- 1192 Quinlan, A. R. & Hall, I. M. (2010). BEDTools: a flexible suite of utilities for comparing
genomic features, *Bioinformatics*, 26, 841-842.
1194 <https://doi.org/10.1093/bioinformatics/btq033>
- 1196 Ringbauer, H., Kolesnikov, A., Field, D. L. & Barton, N. H. (2018). Estimating barriers to
gene flow from distorted isolation-by-distance patterns. *Genetics*, 208, 1231-1245.
1198 <https://doi.org/10.1534/genetics.117.300638>
- 1200 Rix, M. G. & Harvey, M. S. (2012). Phylogeny and historical biogeography of ancient
assassin spiders (Araneae: Archaeidae) in the Australian mesic zone: evidence for
1202 Miocene speciation within Tertiary refugia. *Molecular Phylogenetics and Evolution*, 62,
375-396. <https://doi.org/10.1016/j.ympev.2011.10.009>
- 1204 Rodriguez, L. J., Bain, A., Chou, L.-S., Conchou, L., Cruaud, A., Gonzales, R., Hossaert-
McKey, M., Rasplus, J.-Y., Tzeng, H.-Y. & Kjellberg, F. (2017). Diversification and spatial
1206 structuring in the mutualism between *Ficus septica* and its pollinating wasps in insular
South East Asia. *BMC Evolutionary Biology*, 17, 207. <https://doi.org/10.1186/s12862-017-1034-8>
1208
- 1210 Romiguier, J., Gayral, P., Ballenghien, M., Bernard, A., Cahais, V., Chenuil, A., Chiari, Y.,
Dernat, R., Duret, L., Faivre, N., Loire, E., Lourenco, J. M., Nabholz, B., Roux, C.,
1212 Tsagkogeorga, G., Weber, A. A.-T., Weinert, L. A., Belkhir, K., Bierne, N. Glémin S. &
Galtier, N. (2014). Comparative population genomics in animals uncovers the
1214 determinants of genetic diversity. *Nature*, 515, 261–263.
<https://doi.org/10.1038/nature13685>
- 1216 Rønsted, N., Weiblen, G. D., Savolainen, V. & Cook, J. M. (2008). Phylogeny,
1218 biogeography, and ecology of *Ficus* section Malvanthera (Moraceae). *Molecular
Phylogenetics and Evolution*, 48, 12-22. <https://doi.org/10.1016/j.ympev.2008.04.005>
- 1220 Schiffer, M., Kennington, W. J., Hoffmann, A. A. & Blacket, M. J. (2007). Lack of genetic
1222 structure among ecologically adapted populations of an Australian rainforest *Drosophila*
species as indicated by microsatellite markers and mitochondrial DNA sequences.
1224 *Molecular Ecology*, 16, 1687-700. . <https://doi.org/10.1111/j.1365-294X.2006.03200.x>

- 1226 Schäuble, C. S., & Moritz, C. (2001). Comparative phylogeography of two open forest
1228 frogs from eastern Australia. *Biological Journal of the Linnean Society*, 74, 157-70.
<https://doi.org/10.1111/j.1095-8312.2001.tb01384.x>
- 1230 Schneider, C. J., Cunningham, M. & Moritz, C. (1998). Comparative phylogeography and
1232 the history of endemic vertebrates in the Wet Tropics rainforests of Australia. *Molecular
Ecology*, 7, 487-498. <https://doi.org/10.1046/j.1365-294x.1998.00334.x>
- 1234 Segar, S. T., Dunn, D. W., Darwell, C. T. & Cook, J. M. (2014). How to be a fig wasp down
1236 under: the diversity and structure of an Australian fig wasp community. *Acta Oecologica*,
57, 17-27. <https://doi.org/10.1016/j.actao.2013.03.014>
- 1238 Siddall, M., Rohling, E. J., Thompson, W. G. & Waelbroeck, C. (2008). Marine isotope
1240 stage 3 sea level fluctuations: data synthesis and new outlook. *Reviews of Geophysics*,
46. <https://doi.org/10.1029/2007RG000226>
- 1242 Simao, F. A., Waterhouse, R. M., Ioannidis, P., Kriventseva, E. V. & Zdobnov, E. M.
1244 (2015). BUSCO: assessing genome assembly and annotation completeness with single-
copy orthologs. *Bioinformatics*, 31, 3210-3212.
<https://doi.org/10.1093/bioinformatics/btv351>
- 1246 Smissen, P. J., Melville, J., Sumner, J. & Jessop, T. S. (2013). Mountain barriers and river
1248 conduits: phylogeographical structure in a large, mobile lizard (Varanidae: *Varanus varius*)
from eastern Australia. *Journal of Biogeography*, 40, 1729–1740.
1250 <https://doi.org/10.1111/jbi.12128>
- 1252 Smit, A. F. A., Hubley, R. & Green, P. (2013-2015). *RepeatMasker Open-4.0*. 2013-2015
1254 <<http://www.repeatmasker.org>>.
- 1256 Sousa, V. & Hey, J. (2013). Understanding the origin of species with genome-scale data:
1258 modelling gene flow. *Nature Reviews Genetics*, 14, 404–414.
<https://doi.org/10.1038/nrg3446>
- 1260 Stone, G. N., White, S. C., Csóka, G., Melika, G., Mutun, S., Péntzes, Z., Sadeghi, S. E.,
1262 Schönrogge, K., Tavakoli, M. & Nicholls, J. A. (2017). Tournament ABC analysis of the
western palaeartic population history of an oak gallwasp, *Synergus umbraculus*.
Molecular Ecology, 26, 6685-6703. doi: 10.1111/mec.14372
- 1264 Strasburg, J. L., & Rieseberg, L. H. (2010). How robust are “Isolation with Migration”
analyses to violations of the IM model? A simulation Study. *Molecular Biology and
1266 Evolution*, 27, 297-310. <https://doi.org/10.1093/molbev/msp233>
- 1268 Stuart-Fox, D. M., Schneider, C. J., Moritz, C. & Couper, P. J. (2001). Comparative
1270 phylogeography of three rainforest restricted lizards from mid-east Queensland. *Australian
Journal of Zoology*, 49, 119–127. <https://doi.org/10.1071/ZO00092>
- 1272 Stumpf, M. P. & McVean, G. A. (2003). Estimating recombination rates from population-
genetic data. *Nature Reviews Genetics*, 4, 959-68. <https://doi.org/10.1038/nrg1227>
- 1274 Sutton, T.L., DeGabriel, J. L., Riegler, M. & Cook, J.M. (2018). A temperate pollinator with
1276 high thermal tolerance is still susceptible to heat events predicted under future climate
change. *Ecological Entomology*, 43, 506-512. <https://doi.org/10.1111/een.12528>

1278 Sutton, T. L., Riegler, M. & Cook, J. M. (2016). One step ahead: a parasitoid disperses
1280 farther and forms a wider geographic population than its fig wasp host. *Molecular Ecology*,
25, 882-94. <https://doi.org/10.1111/mec.13445>

1282 Tian, E., Nason, J. D., Machado, C. A., Zheng, L., Yu, H. & Kjellberg, F. (2015). Lack of
1284 genetic isolation by distance, similar genetic structuring but different demographic histories
in a fig-pollinating wasp mutualism. *Molecular Ecology*, 24, 5976–5991.
<https://doi.org/10.1111/mec.13438>

1286 Treangen, T. J. & Salzberg, S. L. (2011). Repetitive DNA and next-generation sequencing:
1288 computational challenges and solutions. *Nature Reviews Genetics*, 13, 36-46.
<https://doi.org/10.1038/nrg3117>

1290 Van der Auwera, G. A., Carneiro, M. O., Hartl, C., Poplin, R., Del Angel, G., Levy-
Moonshine, A., Jordan, T., Shakir, K., Roazen, D., Thibault, J., Banks, E., Garimella, K.,
1292 Altshuler, D., Gabriel, S. & DePristo M, (2013). From FastQ data to high confidence variant
calls: the Genome Analysis Toolkit best practices pipeline. *Current Protocols in*
1294 *Bioinformatics*, 43, 11.10.1-33. <https://doi.org/10.1002/0471250953.bi1110s43>.

1296 Van Meerbeeck, C. J., Renssen, H. & Roche, D. M. (2009). How did Marine Isotope Stage
3 and Last Glacial Maximum climates differ? – Perspectives from equilibrium simulations.
1298 *Climate of the Past*, 5, 33-51. <https://doi.org/10.5194/cp-5-33-2009>

1300 Wachi, N., Kusumi, J., Tzeng, H. Y. & Su, Z. H. (2016). Genome-wide sequence data
suggest the possibility of pollinator sharing by host shift in dioecious figs (Moraceae,
1302 *Ficus*). *Molecular Ecology*, 25, 5732-5746. <https://doi.org/10.1111/mec.13876>

1304 Wall, J. D. (2003). Estimating ancestral population sizes and divergence times. *Genetics*,
163, 395-404.

1306 Walton, W., Stone, G. N. & Lohse, K. (2020). Diverse demographic histories in a guild of
1308 hymenopteran parasitoids. *Molecular Ecology*, (in review). Add information or delete as
required.

1310 Wang, Z. & Liu, K. J. (2016). A performance study of the impact of recombination on
1312 species tree analysis. *BMC Genomics*, 17, 785. [https://doi.org/10.1186/s12864-016-3104-](https://doi.org/10.1186/s12864-016-3104-5)
1314 5

1316 Ware, A. B. & Compton, S. G. (1994)a. Dispersal of adult female fig wasps. 1. Arrivals and
departures. *Entomologia Experimentalis et Applicata*, 73, 221–229.
<https://doi.org/10.1111/j.1570-7458.1994.tb01859.x>

1318 Ware, A. B. & Compton, S. G. (1994)b. Dispersal of adult female fig wasps. 2. Movements
1320 between trees. *Entomologia Experimentalis et Applicata*, 73, 231–238.
<https://doi.org/10.1111/j.1570-7458.1994.tb01860.x>

1322 Weber, L. C., Van der Wal, J., Schmidt, S., McDonald, W. J. F., Shoo, L. P. & Ladiges, P.
1324 (2014). Patterns of rain forest plant endemism in subtropical Australia relate to stable
mesic refugia and species dispersal limitations. *Journal of Biogeography*, 41, 222-38.
1326 <https://doi.org/10.1111/jbi.12219>

1328 Werren, J. H., Richards, S., Desjardins, C. A., Niehuis, O., Gadau, J. & Colbourne, J. K.
 1330 (2010). Functional and evolutionary insights from the genomes of three parasitoid *Nasonia*
 species. *Science*, 327, 343-348. <https://doi.org/10.1126/science.1178028>

1332 Xiao, J. H., Yue, Z., Jia, L. Y., Yang, X. H., Niu, L. H., Wang, Z., ... Huang, D. W. (2013).
 1334 Obligate mutualism within a host drives the extreme specialization of a fig wasp genome.
Genome Biology, 14, R141. <https://doi.org/10.1186/gb-2013-14-12-r141>

1336 Yang, L. Y., Machado, C. A., Dang, X. D., Peng, Y. Q., Yang, D. R., Zhang, D. Y. & Liao,
 1338 W. J. (2015). The incidence and pattern of co-pollinator diversification in dioecious and
 monoecious figs. *Evolution*, 69, 294-304. <https://doi.org/10.1111/evo.12584>

1340 Yu, H., Tian, E. W., Zheng, L. N., Deng, X. X., Cheng, Y. F., Chen, L. F. ... & Kjellberg, F.
 1342 (2019) Multiple parapatric pollinators have radiated across a continental fig tree displaying
 clinal genetic variation. *Molecular Ecology*, 28, 2391-2405.
<https://doi.org/10.1111/mec.15046>

1344 Zerbino, D. R. & Birney, E. (2008). Velvet: algorithms for *de novo* short read assembly
 1346 using de Bruijn graphs. *Genome Research*, 18, 821-9.
<https://doi.org/10.1101/gr.074492.107>

1348 Zhang, J., Kobert, K., Flouri, T. & Stamatakis, A. (2014). PEAR: a fast and accurate
 1350 Illumina Paired-End reAd mergeR. *Bioinformatics*, 30, 614-620.
<https://doi.org/10.1093/bioinformatics/btt593>

1352

1354 **Data Accessibility**

The short read data and the *SPAdes* genome assembly have been deposited at the ENA
 1356 short read archive (number PRJEB35527).

CO1 barcode sequences for the four *Pleistodontes nigriceps* individuals have been
 1358 deposited in Genbank: N1 MF597824; N2 MF597800; S1 MF597825; S2 MF597826).

A *Mathematica* notebook and blockwise sequence data for this study are available from
 1360 the Dryad repository, doi: (added on acceptance).

1362 **Author contributions.**

LC, GNS, KL and JMC designed the research. LC, LB, KL and JH performed the research
 1364 and analysed the data. GNS, KL and LC wrote the paper, with editorial input from all
 authors.

1366

1368

1370

1372

1374

1376

Table 1. Maximum composite likelihood estimates (MCLE) of model support and demographic parameters under (a) IM and (b) ADM models. The best supported model is highlighted in bold with parameter 95% confidence intervals estimated by parametric bootstrap. $n(N_e)$ indicates the number of population size parameters in the model, and N_a indicates the population(s) retaining the ancestral population size. Mig indicates the direction of gene flow in the model, with 0 indicating a model with no gene flow. Model support is measured relative to the best fit model in each class (IM9 in (a) and ADM10 in (b)). Likelihood ratio tests (LRT) were calculated as $2\Delta\ln L$.

(a) IM (Isolation with migration) models

Model	$n(N_e)$	N_a	Mig	$\Delta\ln L$	LRT	T	B	M
IM1	1	both	0	-22,877	n/a	3.938	n/a	n/a
IM2	1	both	N->S	-5,078	35,598	5.210	n/a	0.044
IM3	1	both	S->N	-1,122	43,509	5.533	n/a	0.057
IM4	2	N	0	-22,740	274	3.523	1.198	n/a
IM5	2	S	0	-21,298	3,157	3.808	1.055	n/a
IM6	2	N	N->S	-2,855	36,885	6.417	0.774	0.051
IM7	2	N	S->N	-500	44,479	6.193	0.872	0.061
IM8	2	S	N->S	-1,757	39,081	4.573	1.348	0.065
IM9	2	S	S->N	0	45,481	5.137 (4.784-5.489)	1.193 (1.089-1.296)	0.071 (0.045-0.097)
IM9 sim						0.299	5.150	1.197

(b) ADM (Instantaneous admixture) models

Model	$n(N_e)$	N_a	Mig	$\Delta\ln L$	LRT	T_{adm}	T	B	f
ADM1	1	both	0	-6,349	37,230	n/a	4.969	n/a	0.015
ADM2	1	both	0	-6,349	37,230	n/a	4.969	n/a	0.015
ADM3	1	both	N->S	-4,995	2,634	2.194	6.122	n/a	0.296
ADM4	1	both	S->N	-1,110	10,478	1.970	6.095	n/a	0.257
ADM5	2	N	0	-6,536	36,583	n/a	4.193	1.313	0.014
ADM6	2	S	0	-5,632	35,508	n/a	4.631	1.152	0.017
ADM7	2	N	N->S	-1,645	9,784	1.642	5.195	1.329	0.257
ADM8	2	N	S->N	-172	10,919	2.003	6.872	0.857	0.256
ADM9	2	S	N->S	-1975	9,122	2.281	7.276	0.767	0.291
ADM10	2	S	S->N	0	11,263	1.656 (1.509-1.802)	5.643 (5.317-5.966)	1.181 (1.093-1.269)	0.239 (0.206-0.272)
ADM10 sim						0.302	1.653	5.604	1.184

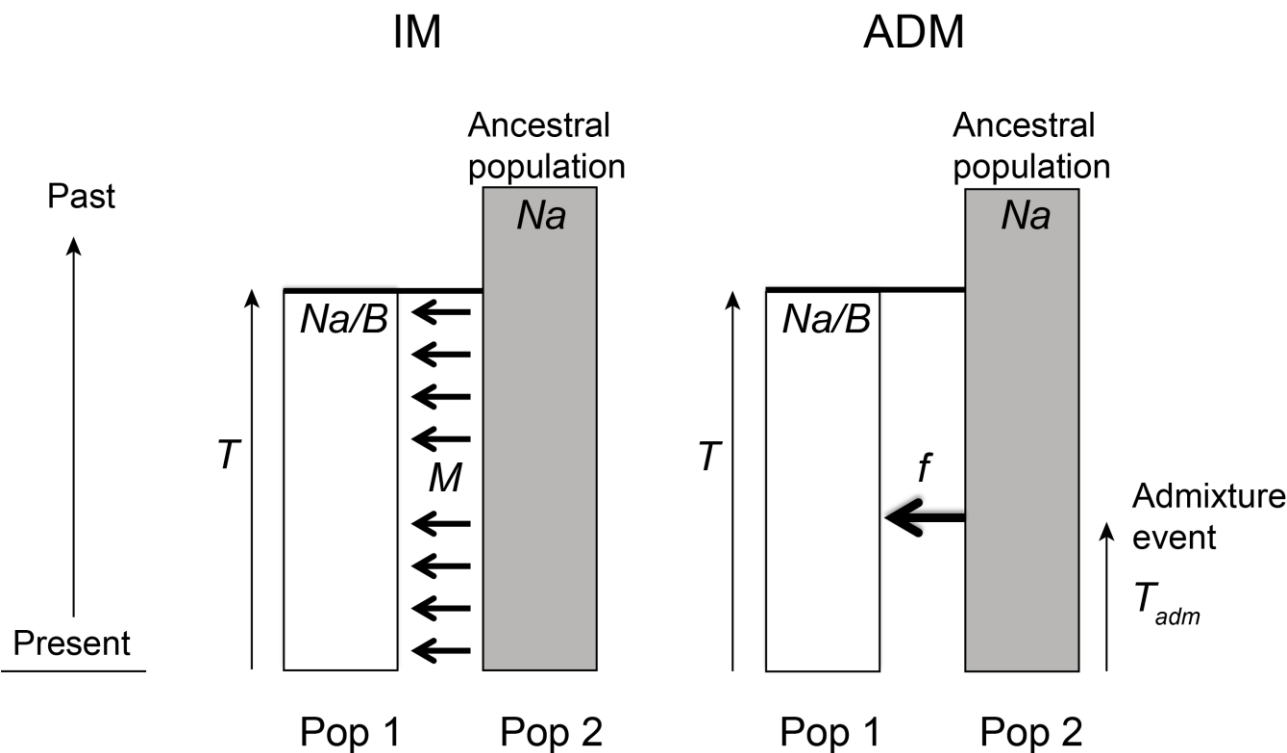


Figure 1. The IM (divergence with continuous migration) and ADM (divergence with instantaneous admixture) models of population divergence with gene flow, showing the demographic parameters estimated in our blockwise method analyses. Gene flow can be modelled in either direction. N_a represents the size of an ancestral population extending back into the past that splits into two daughter populations at time T (scaled by $2N_e$) generations. One population retains the same population size N_a , and one is free to have a new population size N_a/B , where B is a scaling factor. Post divergence gene flow in the IM model is a continuous process with total M per generation = $4N_e * m$, where m is the individual migration rate per generation. In the ADM model, gene flow is modelled as an instantaneous admixture event at time T_{adm} (scaled by $2N_e$) generations ago, at which a fraction f of lineages in the source population are transferred to the receiving population.



1392 **Figure 2.** Map of the east coast of Australia, showing the distribution of *Ficus watkinsiana*
 1394 (*green*) (after Dixon (2003)), the Burdekin and St. Lawrence Gaps (*black*), and the four
 1396 sample sites (*blue circles*). The two Northern sampling sites are Kamerunga (1) and Kairi
 (2); the two Southern sites are Settlers Rise (3) and Main Range (4). The location of the
 study region within the whole of Australia is shown by the box in the inset at bottom left.

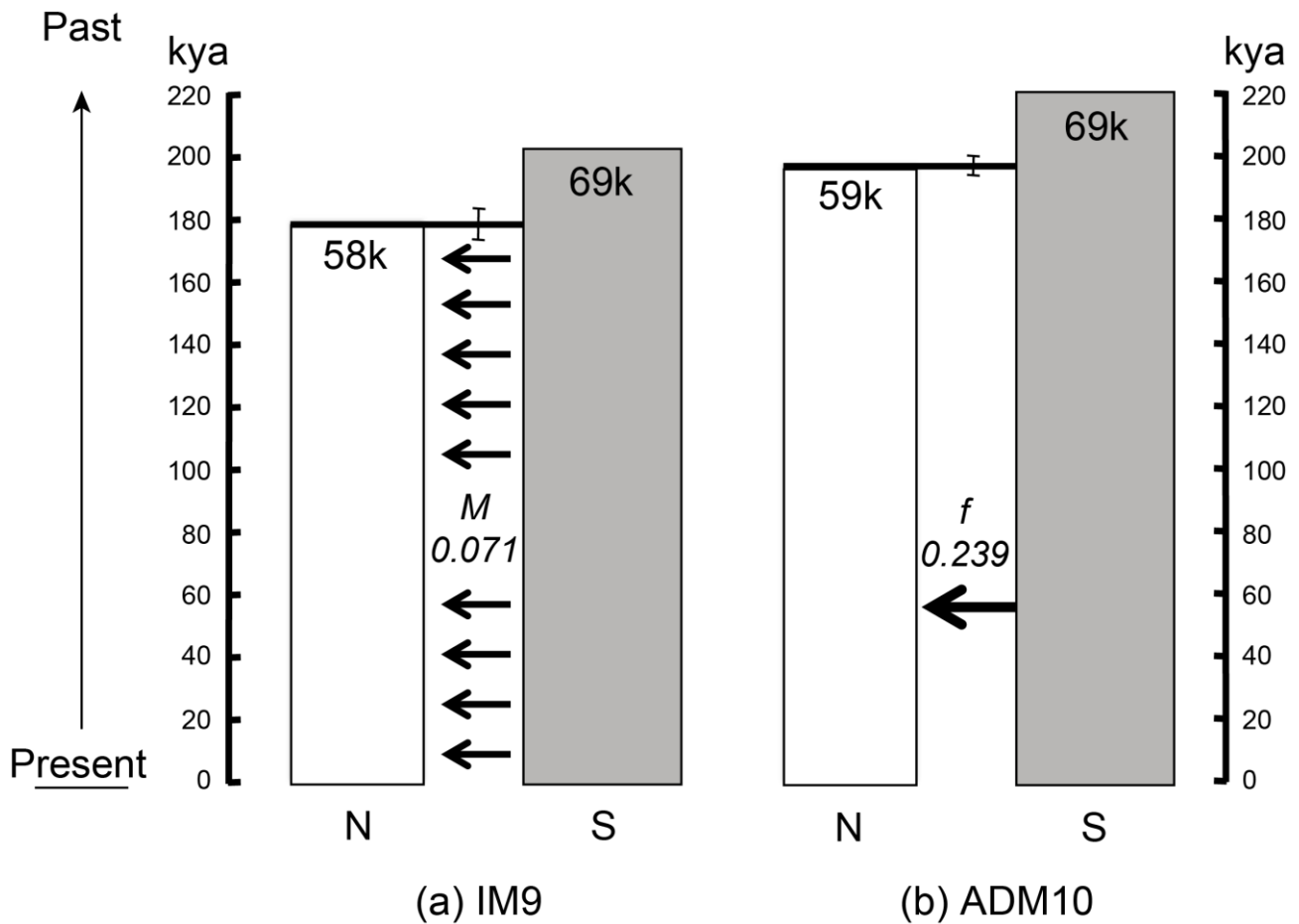


Figure 3. The best supported models under (a) IM and (b) ADM scenarios for *Pleistodontes nigriventris*. Bars joining the populations at divergence are the 95% confidence intervals for the divergence time. Time is measured in thousands of years, assuming 4 generations per year. The population widths are scaled according to their population size estimates (given in thousands). 95% confidence intervals for other parameters are given in Table 1.

